Attachment 8 Aquatic Invasive Species Discipline Report



Aquatic Invasive Species Discipline Report

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Executive Summary

This Aquatic Invasive Species Discipline Report describes the potential impacts of the Capitol Lake – Deschutes Estuary Long-Term Management Project on aquatic invasive species in the area surrounding the project. The Capitol Lake – Deschutes Estuary includes the 260-acre Capitol Lake Basin, located on the Washington State Capitol Campus, in Olympia, Washington. Long-term management strategies and actions are needed to address issues in the Capitol Lake – Deschutes Estuary project area. An Environmental Impact Statement (EIS) is being prepared to document the potential environmental impacts of various alternatives and determine how these alternatives meet the long-term objectives identified for the watershed.

Aquatic invasive species (AIS) include nonnative plants and animals that rely on the aquatic environment for a portion of their life cycle and can spread to new areas of the state, causing economic or environmental harm. The impacts of construction and operation of each alternative are assessed based on the potential of project alternatives to result in changes in abundance or distribution of AIS within or outside the project area from AIS transport into or out of the project area. Where impacts are identified, the report discusses measures that can be taken to minimize or mitigate potential impacts. The analysis examines the No Action Alternative, as well as three build alternatives: Managed Lake, Estuary, and Hybrid.

Under the No Action Alternative, Capitol Lake would remain closed to the public and AIS would continue to be managed from containment and using methods aimed at maintaining low population densities. The New Zealand mudsnail population is not likely to substantially increase within the lake or move far outside the lake because it appears not to have done so in at least 10 years since its introduction in 2009, based on the most recent survey in 2015 and lack of reported sightings outside the lake since then. The containment of these AIS is a highly effective method of avoiding the spread because of the absence of public access, which is the primary way that AIS are spread to new waterbodies. In the absence of public access or meaningful intervention, the populations of other AIS invertebrates, fish, waterfowl, and mammals would be expected to continue to expand within the basin at current low rates. Based on this, under the No Action Alternative, the risk for AIS in Capitol Lake to spread to otherwise non-invaded water bodies is expected to be **less-than-significant**.

Prior to construction of all action alternatives, Capitol Lake would be treated to significantly reduce AIS populations within the waterbody and reduce the risk of potential spread once construction activities began. Management actions prescribed under an AIS Management Plan may include the use of herbicides/pesticides to reduce the number of purple loosestrife seeds and live New Zealand mudsnails. Under all build alternatives, areas of Capitol Lake would be dredged during construction and the dredge sediment would be beneficially reused within the Capitol Lake Basin to create habitat areas. Reusing the dredged material onsite is a key design feature that would further minimize the potential transport of AIS outside the project area. Off-site transport of excess dredge material for the Estuary and Hybrid Alternatives would require additional Best Management Practices (BMPs) or treatment to prevent the spread of AIS in those materials.

The Estuary and Hybrid Alternatives include removing the 5th Avenue Dam, reintroducing saltwater to the basin and creating a brackish environment. The introduction of saltwater would have a **substantial beneficial effect** on freshwater AIS populations that are intolerant to higher salinities.. However, despite early treatment of the Capitol lake Basin, there will remain a potential for AIS populations of purple loosestrife and New Zealand mudsnails to spread outside the project area following the 5th Avenue Dam removal given their tolerance of higher salinities. Because of this potential, construction could have a **significant impact** on AIS distribution under the Estuary and Hybrid Alternatives, compared to **less-than-significant** impact for the Managed Lake Alternative.

Under all action alternatives, the basin would be reopened to the public and would feature boardwalks and dock and the waters would be open for non-motorized recreational vessels. The increase in pedestrian and boating activity could increase the spread of AIS outside the project area, as well as potentially introduce new AIS to the project area. A critical component of all action alternatives is the installation and required use of decontamination stations for all vessels or other recreational gear that enters and exits the waterbody. All action alternatives would also avoid or minimize spread of AIS by posting educational signage warning recreational users of AIS presence and AIS populations would be monitored as prescribed in an AIS Management Plan to ensure decontamination effectiveness.

Construction and operation impacts of the No Action and Build Alternatives are summarized in Tables ES.1 and ES.2.

Impact	Impact Finding	Mitigation (Summarized)	Significant and Unavoidable Adverse Impact?
Managed Lake Alternative			
Increase spread of aquatic invasive plants and animals	Less-than- significant	In addition to BMPs and other measures included in Section 5.7.1.2:	No
		 Dredge material would not be removed from the project area. 	
Estuary Alternative			
Increase spread of aquatic invasive plants and animals	Significant	 In addition to BMPs and other measures included in Section 5.7.1.2: Implement measures outlined in an AIS Management Plan to eradicate or reduce the purple loosestrife population and reduce the New Zealand mudsnail population prior to dam removal. 	No
Hybrid Alternative			
Increase spread of aquatic invasive plants and animals	Significant	In addition to BMPs and other measures included in Section 5.7.1.2:	No
		 Implemented measures outlined in an AIS Management Plan to eradicate or reduce the purple loosestrife population and reduce the New Zealand mudsnail population prior to dam removal. 	

Table ES.1 Summary of Construction Impacts and Mitigation Measures

Impact	Impact Finding	Mitigation (Summarized)	Significant and Unavoidable Adverse Impact?
No Action Alternative			
Increase spread of aquatic invasive plants and animals	Less-than-significant	Containment and other minor management practices would continue to reduce potential spread outside the project area.	No
Managed Lake Alternative			
Increase spread of aquatic invasive plants and animals	Less-than-significant	 In addition to BMPs and other measures included in Section 5.7.1.3: Periodic treatment to reduce the New Zealand mudsnail population according to AIS Management Plan. 	No
Estuary Alternative			
Increase spread of aquatic invasive plants and animals	Less-than-significant	BMPs and other measures to minimize impacts are included in Section 5.7.1.3.	No
	Substantial Beneficial Effects (related to decreased distribution and abundance of AIS species that are intolerant to saltwater)		
Hybrid Alternative			
Increase spread of aquatic invasive plants and animals	Less-than-significant	BMPs and other measures to minimize impacts are included in Section 5.7.1.3.	No
	Substantial Beneficial Effects (related to decreased distribution and abundance of AIS species that are intolerant to saltwater)		

Table ES.2 Summary of Operations Impacts (including Benefits) and Mitigation Measures



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List of Acronyms and Abbreviations

Acronyms/ Abbreviations	Definition
AIS	Aquatic invasive species
EIS	Environmental Impact Statement
Enterprise Services	Washington State Department of Enterprise Services
GA	Washington State Department of General Administration
mm	Millimeters
NAM	Northwest Aquatics Management, LLC
WDFW	Washington Department of Fish and Wildlife



1.0 Introduction and Project Description

1.1 **PROJECT DESCRIPTION**

The Capitol Lake – Deschutes Estuary includes the 260-acre Capitol Lake Basin, located on the Washington State Capitol Campus, in Olympia, Washington. The waterbody has long been a valued community amenity. Capitol Lake was formed in 1951 following construction of a dam and provided an important recreational resource. Historically, the Deschutes Estuary was used by local tribes for subsistence and ceremonial purposes. Today, the expansive waterbody is closed to active public use. There are a number of environmental issues including the presence of invasive species, exceedances of water quality (WQ) standards, and inadequate sediment management.

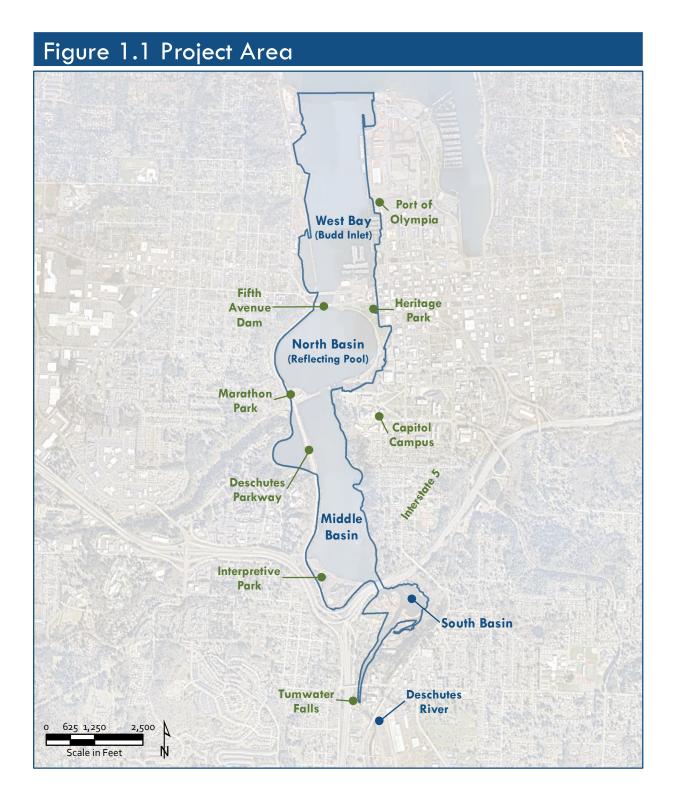
The Washington State Department of Enterprise Services (Enterprise Services) is responsible for the stewardship, preservation, operation, and maintenance of the Capitol Lake Basin. The 260-acre Capitol Lake Basin is maintained by Enterprise Services under long-term lease agreement from the Washington Department of Natural Resources.

In 2016, as part of Phase 1 of long-term planning, a diverse group of stakeholders, in collaboration with the state, identified shared goals for long-term management and agreed an Environmental Impact Statement (EIS) was needed to evaluate a range of alternatives and identify a preferred alternative. In 2018, the state began the EIS process. The EIS evaluates four alternatives, including a Managed Lake, Estuary, Hybrid, and a No Action Alternative.

The long-term management alternatives are evaluated against the shared project goals of: improving water quality; managing sediment accumulation and future deposition; improving ecological functions; and enhancing community use of the resource. Refer to Figure 1.1 for the project area for long-term management. The Final EIS will identify a preferred environmentally and economically sustainable long-term management alternative for the Capitol Lake – Deschutes Estuary.

The EIS process maintains engagement with the existing Work Groups, which include the local governments, resource agencies, and tribe. It also provides for expanded engagement opportunities for the public, such as a community sounding board.

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1.2 SUMMARY OF PROJECT ALTERNATIVES

1.2.1 Managed Lake Alternative

The Managed Lake Alternative would retain the 5th Avenue Dam in its existing configuration. The 5th Avenue Dam would be overhauled to significantly extend the serviceable life of the structure. The reflecting pool within the North Basin would be maintained, and active recreational use would be restored in this area. Sediment would be managed through initial construction dredging and recurring maintenance dredging in the North Basin only. Sediment from construction dredging would be used to create habitat areas in the Middle Basin to support improved ecological function, habitat complexity, and diversity. Sediment would continue to accumulate and over time would promote a transition to freshwater wetlands in the South and Middle Basins. Boardwalks, a 5th Avenue Pedestrian Bridge, a dock, and a boat launch would be constructed for community use.

If selected as the Preferred Alternative, adaptive management plans would be developed to maintain water quality, improve ecological functions, and manage invasive species during the design and permitting process.

1.2.2 Estuary Alternative

Under the Estuary Alternative, the 5th Avenue Dam would be removed, and an approximately 500-footwide (150-meter-wide) opening would be established in its place. This would reintroduce tidal hydrology to the Capitol Lake Basin, returning the area to estuarine conditions where saltwater from Budd Inlet would mix with freshwater from the Deschutes River. Sediment would be managed through initial construction dredging in the Capitol Lake Basin and recurring maintenance dredging within West Bay. Dredged materials from construction dredging would be used to create habitat areas in the Middle and North Basins to promote ecological diversity, though tideflats would be the predominant habitat type. Boardwalks, a 5th Avenue Pedestrian Bridge, a dock, and a boat launch would be constructed for community use. This alternative also includes stabilization along the entire length of Deschutes Parkway to avoid undercutting or destabilization from the tidal flow. Existing utilities and other infrastructure would be upgraded and/or protected from reintroduced tidal hydrology and saltwater conditions.

If selected as the Preferred Alternative, adaptive management plans would be developed to improve ecological functions and manage invasive species during the design and permitting process.

1.2.3 Hybrid Alternative

Under the Hybrid Alternative, the 5th Avenue Dam would be removed, and an approximately 500-footwide (150-meter-wide) opening would be established in its place. Tidal hydrology would be reintroduced to the western portion of the North Basin and to the Middle and South Basins. Within the North Basin, a curved and approximately 2,600-foot-long (790-meter-long) barrier wall with a walkway would be constructed to create an approximately 45-acre saltwater reflecting pool adjacent to Heritage Park. A freshwater (groundwater-fed) reflecting pool was also evaluated for this EIS. Construction and maintenance of this smaller reflecting pool, in addition to restored estuarine conditions in part of the Capitol Lake Basin, gives this alternative its classification as a hybrid. Sediment would be managed through initial construction dredging in the Capitol Lake Basin and recurring maintenance dredging within West Bay. In the Middle and North Basins, constructed habitat areas would promote ecological diversity, though tideflats would be the predominant habitat type. Boardwalks, a 5th Avenue Pedestrian Bridge, a dock, and a boat launch would be constructed for community use. This alternative also includes stabilization along the entire length of Deschutes Parkway to avoid scour or destabilization. Existing utilities and other infrastructure would be upgraded and/or protected from reintroduced tidal hydrology and saltwater conditions.

If selected as the Preferred Alternative, adaptive management plans would be developed before operation of the alternative to improve ecological functions and manage invasive species during the design and permitting process. Adaptive management would also be needed for a freshwater reflecting pool, but not for a saltwater reflecting pool.

1.2.4 No Action Alternative

The No Action Alternative represents the most likely future expected in the absence of implementing a long-term management project. The No Action Alternative would persist if a Preferred Alternative is not identified and/or if funding is not acquired to implement the Preferred Alternative. A No Action Alternative is a required element in a SEPA EIS and provides a baseline against which the impacts of the action alternatives (Managed Lake, Estuary, Hybrid) can be evaluated and compared.

The No Action Alternative would retain the 5th Avenue Dam in its current configuration, with limited repair and maintenance activities, consistent with the scope and scale of those that have received funding and environmental approvals over the past 30 years. In the last 30 years, the repair and maintenance activities have been limited to emergency or high-priority actions, which occur sporadically as a result of need and funding appropriations.

Although Enterprise Services would not implement a long-term management project, current management activities and ongoing projects in the Capitol Lake Basin would continue. Enterprise Services would continue to implement limited nuisance and invasive species management strategies.

In the absence of a long-term management project, it is unlikely that Enterprise Services would be able to procure funding and approvals to manage sediment, improve water quality, improve ecological functions, or enhance community use. The No Action Alternative does not achieve the project goals.

1.3 CONSTRUCTION METHODS FOR THE ACTION ALTERNATIVES

This impact analysis relies upon the construction method and anticipated duration for the action alternatives, which are described in detail in Chapter 2 of the EIS.



2.0 Regulatory Context

2.1 **RESOURCE DESCRIPTION**

This report describes aquatic invasive plant and animal species in the study area and evaluates the potential impact of each project alternative on invasive species populations and their spread outside the study area. Invasive species are nonnative organisms that cause economic or environmental harm and can spread to new areas of the state (RCW 79A.25.310). Aquatic invasive species (AIS) are invasive species that rely on the aquatic environment for all or a portion of their life cycle. Nonnative, non-indigenous, or exotic species are those species that are present outside of their natural range. Unlike invasive species, nonnative species do not necessarily hinder or prevent the survival of native species within the ecosystem. Nuisance species may be native or nonnative and may cause ecological and economic harm.

2.2 FEDERAL AND STATE LAWS, PLANS, AND POLICIES

Several federal, state, and local government policies, regulations, and ordinances relating to the eradication and management of invasive species apply to this project. Such regulations and policies influence planning, land use, and management activities that can impact fish and wildlife species and their habitats within the study area. Tables 2-1 and 2-2 summarize federal and state regulations and programs for AIS eradication and management. While several federal laws may address AIS, such as the Clean Water Act, Coastal Management Zone, Fish and Wildlife Conservation Act, and the Magnuson-Stevens Fishery Conservation and Management Act, the primary purpose of these regulations is not invasive species management. For more information on these regulations, see the Fish and Wildlife Discipline Report and the Wetlands and Vegetation Discipline Report.

Regulatory Program or Policies	Lead Agency	Description
Aquatic Nuisance Species Task Force	U.S. Fish and Wildlife Service	U.S. Fish and Wildlife Service coordinates with federal agencies and partners to prevent and control AIS and provide expertise and support for action-oriented partnerships (84 FR 56832; October 23, 2019).
Executive Order 13112 – Invasive Species	National Invasive Species Council	This executive order prevents federal agencies from authorizing, funding, or carrying out actions that are likely to cause or promote the introduction or spread of invasive species (USDA 1999).
Federal Insecticide, Fungicide, and Rodenticide Act	U.S. Environmental Protection Agency	This Act requires the study of the consequences of pesticide usage and requires users to register when purchasing pesticides. All pesticides must be registered (licensed) by USEPA. Registration ensures that pesticides will be properly labeled and used in accordance with specifications that will not cause unreasonable harm to the environment (7 USC §136 et seq. 1996).
National Invasive Species Act	U.S. Department of Agriculture	An Act to prevent and control infestations of the coastal inland waters of the United States by the zebra mussel and other nonindigenous aquatic nuisance species (H.R. 4283; 104th Congress 1996).
Executive Order 13751 – Safeguarding the Nation from the Impacts of Invasive Species	National Invasive Species Council	An Act that directs actions to coordinate federal prevention and control efforts related to invasive species (USDA 2016).

Table 2-1. Aquatic Invasive Species Federal Regulatory Programs and Policies.

Table 2-2. Aquatic Invasive Species State Regulatory Programs and Policies.

Regulatory Program or Policies	Lead Agency	Description
Invasive Species Policy Coordination	Washington Invasive Species Council	The council is tasked with providing policy level direction, planning, and coordination to the various public and private entities working throughout Washington State in order to prevent and control the spread of harmful invasive species.
Aquatic Invasive Species Prevention	Washington Department of Fish and Wildlife	Operators of watercraft not registered in Washington State, seaplanes, and commercial transporters of specified vessel types must purchase AIS prevention permits to help prevent the spread of aquatic invasive species in Washington (WDFW 2020a).
Invasive Species Law	Washington Department of Fish and Wildlife	RCW 77.15.805 enforces the Invasive Species Law. RCW 77.135 prohibits transporting any visible native and nonnative aquatic animals, plants, or other organisms, as well as raw water on any boat, trailer, fishing gear, or bait well.

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Regulatory Program or Policies	Lead Agency	Description
Washington State Noxious Weed Law	Washington State Noxious Weed Control Board	RCW 17.10 provides for the creations of the state noxious weed board, county noxious weed boards, the state noxious weed list, landowner responsibilities in noxious weed control, and the ability of employees of the state weed board, county weed boards, or Washington State Department of Agriculture to carry out their duties and authorities assigned under this chapter.

2.3 LOCAL LAWS, PLANS, AND POLICIES

Table 2-3 describes the regulatory programs implemented by the study area communities.

Table 2-3. Local Laws, Plans, and Policies.	Table 2-3.	Local	Laws,	Plans,	and	Policies.
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Regulatory Program or Policies	Lead Agency	Description
Critical Area Development Regulations	Olympia Municipal Code	This Code creates regulations regarding the presence of invasive species in critical areas, riparian buffers, and mitigation/restoration sites (OMC 18.32).
Fish and Wildlife Habitat Conservation Areas	Tumwater Municipal Code	This Code creates regulations regarding the removal of noxious weeds and invasive plants and addresses introduction of invasive and nonnative plant species (TMC 16.32).
Noxious Weed Control	Thurston County Noxious Weed Control Board	The Thurston County Noxious Weed Control Board fulfills the requirements of RCW 17.10 and establishes policies and procedures in accordance with statutes and state regulations, adopts rules and regulations for noxious weed control, and prepares the annual Noxious Weed List (Thurston County 2020a).



3.0 Methodology

3.1 DATA SOURCES AND COLLECTION

An extensive literature review was conducted for this report. Over 75 data sources were reviewed for the AIS evaluation, including management plans, surveys, databases, and research papers.

3.1.1 Management Plans

The management plans reviewed include vegetation management, reports on permit needs, annual reports of aquatic weed treatments, New Zealand mudsnail management options, and recommendations for future and ongoing invasive species treatments. The primary management plans used in the evaluation that provided recent and relevant information include:

- Deschutes River Estuary Restoration Study: Biological Conditions Report (Garono et al. 2006)
- Capitol Lake New Zealand Mudsnail Management Options (WDFW 2016a)
- Capitol Lake Weed Management Services 2018 Annual Report (NAM 2018)

3.1.2 Surveys

Several surveys have been conducted to monitor the presence and distribution of AIS in Capitol Lake, and some surveys were associated with AIS treatment and removal efforts. The primary surveys used in this evaluation that provided recent and relevant information include:

- Capitol Lake Vertebrate and Invertebrate Inventory (Herrera 2004)
- Capitol Lake Weed Management Services, Bottom Screen Report (NAM 2011)
- 2012 Milfoil Weevil Population Survey in Capitol Lake and East Mitigation Pond (EnviroScience 2013)
- 2015 Survey for *Potamopyrgus antipodarum* (New Zealand Mudsnail) Within a Five-Mile Radius of Capitol Lake (Johannes 2015)

3.1.3 Databases

To maintain the most recent evidence of invasive species presence and spread in the area, several databases were reviewed. The main databases used in this evaluation that provided recent and relevant information include:

- Federal Noxious Weeds (USDA 2020)
- Noxious Weed layer on the Show Me Everything map (Thurston County 2020b)
- Nonindigenous Aquatic Species (USGS 2020)
- Aquatic Invasive Species (WDFW 2020b)
- Washington Marine Vegetation Atlas (WDNR 2020)

3.1.4 Research Papers

Research papers and studies used in this evaluation explored a diverse range of topics including detection, species biology (specifically as it relates to the overall success of an AIS in a nonnative area), population fluctuations, transport and spread, and treatment options and effectiveness. Some examples of research papers used in this evaluation include:

- A review of salinity tolerances for the New Zealand mudsnail (*Potamopyrgus antipodarum*, Gray 1843) and the effect of a controlled saltwater backflush on their survival in an impounded freshwater lake (LeCLair and Cheng 2011)
- Environmental DNA as a new method for early detection of New Zealand mudsnails (Goldberg et al. 2013) A quantitative evaluation of the effect of freezing temperatures on the survival of New Zealand mudsnails (*Potamopyrgus antipodarum*; Gray 1843), in Olympia Washington's Capitol Lake (Cheng and LeClair 2011)
- Survival and passage of ingested New Zealand mudsnails through the intestinal tract of rainbow trout (Bruce et al. 2009)

3.2 SELECTION OF THE STUDY AREA

The study area is based on the local aquatic resources where AIS could be directly affected by the project, and does not include distant water bodies where AIS potentially could be transported to by the project. The study area includes the Capitol Lake Basin (the North Basin, Middle Basin, South Basin, and Percival Cove), Percival Creek up to Highway 101, the Deschutes River upstream of Tumwater Falls, and Lower Budd Inlet extending north from the 5th Avenue Dam to southern end of Priest Point Park near the mouth of Mission Creek (47 o7'N) (Figure 1.1). The study area extends approximately 100 feet from the edge of these waters to encompass emergent plant, amphibian, waterfowl, and mammal AIS.

3.3 ANALYSIS OF IMPACTS

Adverse impacts and beneficial effects related to both construction and long-term operation are evaluated, with a focus on comparatively evaluating the alternatives. In general, construction-related

impacts are primarily associated with lake sediment dredging and dam construction (repair or removal) because that activity represents the major construction impact. Future, long-term adverse impacts, and beneficial effects associated with AIS populations and control actions for each of the four project alternatives are evaluated using a combination of historical trends, current conditions, and future projections of environmental factors affecting AIS.

Assessments of potential adverse impacts and beneficial effects are based on many factors including:

- Anticipated changes in abundance and distribution for each species
- Relative potential for transport and establishment within and outside the study area
- Control priority, eradication potential and potential management options of each species
- Relative effectiveness and non-target species impacts of control measures
- Potential for short- and long-term recreational use restrictions

Qualitative categories such as "less-than-significant" and "significant" are used to assess the relative magnitude of adverse impacts related to AIS. Significant increases in AIS populations or distribution by an alternative are considered to be an adverse impact, whereas significant decreases in AIS populations or distribution are considered a beneficial effect of the alternative.

3.3.1 Identification of Construction Impacts

Sediment dredging, dredged material placement for constructing habitat areas, and dam construction (repair or removal) are the primary construction activities affecting AIS. These activities have the potential adverse impact of spreading existing AIS in Capitol Lake to other water bodies if the following occur:

- AIS associated with suspended sediment and debris care are not contained.
- Construction equipment is not properly decontaminated before it leaves the lake.
- Dredged material is not properly decontaminated before it leaves the lake.
- Equipment is not properly decontaminated before use in the lake, leading to the introduction of new AIS.

Dredging and dredge material placement has the potential beneficial effect of reducing existing AIS populations through burial.

Repairs to the existing 5th Avenue Dam have the potential adverse impact of spreading existing AIS in Capitol Lake to Budd Inlet and adjacent streams if the construction area is not properly contained. Other construction activities potentially affecting AIS by improper equipment decontamination include construction of the boardwalks and docks, and the pedestrian and bicycle bridge.

For this evaluation, the magnitude of short-term adverse impacts is considered less-than-significant or significant, as follows:

- Less-than-significant—Impacts are considered less-than-significant if there would be no substantial increased risk, relative to existing conditions, of introducing new AIS to a water body during construction that would impact native species or recreational use. Any potential short-term effect of construction on the abundance of existing AIS within a water body is considered less-than-significant because AIS are inherently resilient to disturbance and expected to recover from short-term changes in abundance.
- **Significant**—Impacts are considered significant if there would be a substantial increased risk, relative to existing conditions, of introducing new AIS to a water body during construction that would impact native species or recreational use.

The adverse impacts and beneficial effects of construction on each AIS are described for each alternative in Section 5.0 Impacts and Mitigation Measures.

3.3.2 Identification of Operational Impacts

Changes in the salinity and recreational use of the Capitol Lake Basin are the primary operational impacts to AIS that could be positive or negative. That is, increased salinity has the beneficial effect of reducing survivorship of freshwater AIS, and increased recreational use has the potential adverse impact of increasing the potential for further spread of AIS. Maintenance dredging also has the potential for introducing new AIS if equipment is not properly decontaminated or spreading AIS from offsite transport of dredged sediments.

For this evaluation, the magnitude of long-term (operational) adverse impacts are considered lessthan-significant or significant, as follows:

- Less-than-significant—Impacts are considered less-than-significant if no substantial increase (e.g., less than a 50 percent increase) is expected for the abundance or distribution of existing AIS within the study area, or if the potential for existing AIS to spread to other water bodies is expected to not substantially increase and impact native populations or recreational use, relative to the existing condition.
- **Significant**—Impacts are considered significant if a substantial increase (e.g., more than 50 percent increase) is expected for the abundance or distribution of existing AIS within the study area, or if the risk of introducing new AIS would substantially increase for waters inside or outside the study area and impact native populations or recreational use, relative to the existing condition.

Long-term beneficial effects on AIS are evaluated in terms of minor, moderate, or substantial beneficial effects where expected decreases in AIS abundance or distribution inside or outside the study area are considered beneficial. The impact of regular operations on each AIS is described for each alternative in Section 5.0 Impacts and Mitigation Measures.



4.0 Affected Environment

Capitol Lake has a well-documented presence of AIS including plants, invertebrates, fish, waterfowl, and aquatic mammal species. The presence of AIS has resulted in closure of Capitol Lake to all public use. This section briefly describes the species biology, documented presence in Capitol Lake, the ecological impact of the AIS presence, and previous efforts to control the presence and spread of invasive species.

The biology of an invasive species provides information on its ability to outcompete native species for resources. Before an invasive species can become a nuisance in an ecosystem after being introduced, it must survive, grow, and reproduce in the new environmental conditions. Thus, an invasive species must acquire a high rate of population growth, invading new regions before it finally alters the structure and function of the invaded ecosystem (Alonso and Castro-Díez 2008). The existing presence and spread of AIS provide data on the pathway and possible containment. Biological invasions often represent a major ecological and economic problem. The description of ecological impacts will inform the potential impacts or benefits for each alternative. Previous efforts to control AIS presence and spread, both in Capitol Lake and outside the study area, inform future decisions for control efforts that differ based on each alternative explored in later sections.

4.1 AQUATIC INVASIVE PLANTS

Aquatic invasive plants found in Capitol Lake and the surrounding water bodies include purple loosestrife (*Lythrum salicaria*), yellow flag iris (*Iris pseudacorus*), reed canarygrass (*Phalaris arundinacea*), fragrant water lily (*Nymphaea odorata*), Eurasian watermilfoil (*Myriophyllum spicatum*), and curlyleaf pondweed (*Potamogeton crispus*). Table 4-1 describes their classification on the Washington State Noxious Weed List (2020), distribution in the study area, and relative abundance in study area.

Scientific/ Common Name	State Statusª	Water Body	Relative Abundance	Source
Emergent				
<i>Lythrum salicaria</i> Purple loosestrife	Class B High Priority	Capitol Lake, Deschutes River, Budd Inlet	High in the South Basin and Percival Cove; low in other basins	NAM 2018
<i>Iris pseudacorus</i> Yellow flag iris	Class C	Capitol Lake	High in the South and Middle Basins; Moderate in Percival Cove; low in the North Basin.	NAM 2018
<i>Phalaris arundinacea</i> Reed canary grass	Class C	Capitol Lake	Present at unknown locations in 2006	Ecology 2020a
Floating Leaved				
<i>Nymphaea odorata</i> Fragrant waterlily	Class C	Capitol Lake	Moderate in North Basin and Percival Cove; low to zero in Middle Basin and South Basin	NAM 2018
Submersed	•	·		•
<i>Myriophyllum spicatum</i> Eurasian watermilfoil	Class B High Priority	Capitol Lake	Moderate in South Basin, Middle Basin, and Percival Cove; low in North Basin	NAM 2018
Potamogeton crispus Curlyleaf pondweed	Class C	Capitol Lake	Present at unknown locations in 2006 and primarily in the south end of lake in 2004	Ecology 2020a

Table 4-1. Aquatic invasive Plants Observed in the Study Area	Table 4-1.	Aquatic Invasive Plants Observed in the Study Area.
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^a Washington State Noxious Weed Class (WNWCB 2020) or High Priority Invasive Species (WISC 2020c).

Class A noxious weeds are nonnative species whose distribution in Washington is still limited. Eradication of Class A noxious weeds is required by law. Class B noxious weeds are nonnative species whose distribution are limited to portions of Washington. Control of Class B noxious weeds is required in regions where they are not yet widespread in order to prevent new infestations. Class C noxious weeds are widespread in Washington or are of special interest to the agricultural industry (WNWCB 2020).

4.1.1 Presence in Capitol Lake

The biology, ecological and recreational implications, distribution, and abundance of aquatic invasive emergent, floating-leaved, and submerged plants in Capitol Lake are described in the following sections. Distribution, abundance, and management efforts are described in approximate chronological order for each plant species.

4.1.1.1 Emergent Plants

Purple Loosestrife

Purple loosestrife (*Lythrum salicaria*) is a Class B Washington State Listed Noxious Weed (Washington State Noxious Weed Control Board 2020) and a High Priority invasive species (WISC 2020c). Control of

Class B species is required by the Thurston County Noxious Weed Board. The plant is a nonnative emergent species typically found in freshwater and brackish wetlands, along streams, and in other wet areas. It has narrow, lance-shaped leaves; showy purple flowers that occur in erect spikes at the top of stems from late June through October; and a rhizomatous growth pattern. The plant is a vigorous grower that spreads by rhizomes or by seed. Each plant may produce up to 2.7 million seeds annually. The seeds can be viable for several years, but because the seeds are small and carry little food reserves, germination must occur when photosynthesis can occur immediately (USDA 2020). Purple loosestrife forms dense colonies that outcompete native plant species and provide minimal wildlife habitat (Ecology 2001). These dense colonies can also be detrimental to aesthetics and inhibit access to shorelines



Purple Loosestrife (WNWCB 2020)

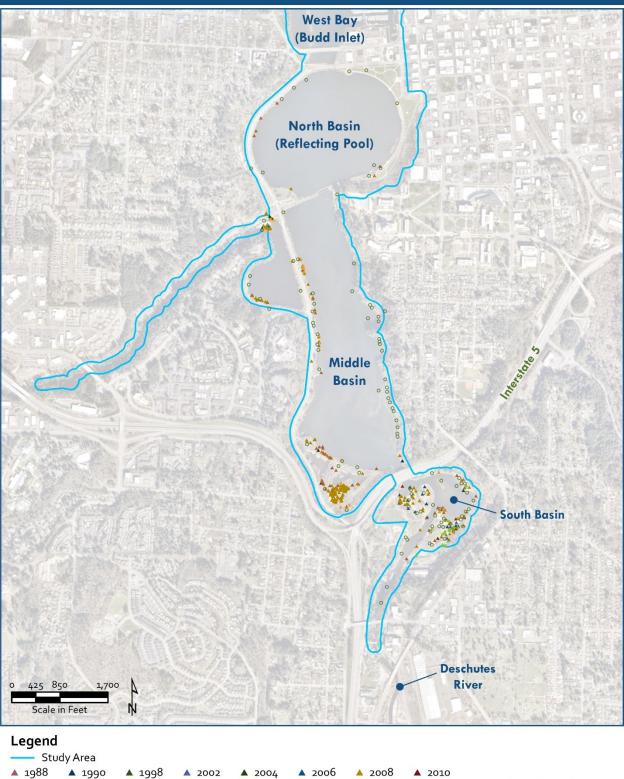
for recreation (RMI 2001). Management of purple loosestrife is a costly effort requiring repeated monitoring and removal efforts to prevent its spread.

Distribution and Abundance Within the Study Area

Purple loosestrife was first discovered in Capitol Lake in 1986. By 1987, the infestation was described as "bad" (Washington State Department of General Administration [GA] 2002). Figure 4-1 presents a map of purple loosestrife historical observations from 1988 through 2010 along with observation points colored red where it had been eradicated. Figure 4-2 presents the most current map of purple loosestrife distribution in 2018 with observation points colored purple (NAM 2018). Purple loosestrife was most abundant in the South Basin but was present along the shorelines of all three basins, the mitigation area, and Percival Cove. Only one plant was observed in the North Basin, and no plants were observed along the east shoreline of the Middle Basin. Although purple loosestrife is a high-priority species based on its aggressive growth and potential impacts on native species, it is not likely significantly impacting native wildlife or recreation in the Capitol Lake Basin based on its current abundance and the emergent plant diversity.

Long-Term Management Project Environmental Impact Statement





▲ 1989

1992

1999

a 2003

▲ 2005

Aquatic Invasive Species Discipline Report

A 2007

a 2009

O Purple Loosestrife Eradicated

Long-Term Management Project Environmental Impact Statement





Legend

Study Area

Purple Loosestrife

Management Approaches

In 1987, the Thurston County Noxious Weed Control Board and the GA began communicating to organize efforts to control purple loosestrife. From 1989 to 1993, various maintenance groups and contracted groups for the GA removed flower heads to prevent seed distribution. From 1993 to 1995, the GA contracted with Resource Management, Inc. to perform aquatic herbicide treatments on purple loosestrife in Capitol Lake, but information on the relative effectiveness of the treatments was not available.

From 1996 to 2000, the GA returned to removing flower heads to control purple loosestrife because the herbicide treatments did not eradicate purple loosestrife. Based on the results of these efforts, it was determined that seed production is controlled by flower head removal, but stem densities still increase in the case of flower head removal. It was determined that future efforts should include treatment of the entire plant along with seed removal. The lack of continuity in treatments from year to year and the lack of monitoring after control efforts were implemented limits the understanding of the efficacy of management actions.

In 1998, wetland soils from the southern end of the Middle Basin were removed and replaced as a mitigation measure for Heritage Park, subsequently soils bordering the constructed wetland mitigation ponds in Interpretive Park were quickly infested with purple loosestrife. In 1999, the Thurston County Noxious Weed Board released 5,000 galerucella beetles (*Galerucella calmariensis*) as a form of biological control for loosestrife populations in the Interpretive Park wetland ponds. Later reports indicate that purple loosestrife showed signs of beetle activity; however, the population of beetles was not dense enough to control the purple loosestrife infestation in this area (NAM 2013, 2017).

In 2000, the infestation of purple loosestrife was determined to have been reduced by 80 percent from levels in 1987 (NAM 2013). In 2001, the GA adopted the Capitol Lake Integrated Purple Loosestrife Management Plan, which established the goal to eradicate purple loosestrife from Capitol Lake and adjacent areas (RMI 2001). This plan describes results from a field survey done by aquatic biologists in December of 2000 and input from GA Grounds Supervisor Mark Robb. According to these sources, approximately 2.5 acres of dense purple loosestrife colonies were present in the wetland mitigation area at the south end of the Middle Basin. Approximately 1.3 acres of moderate-density purple loosestrife were present in the mitigation area and along the western shorelines of the Middle Basin. In December 2000, the total area of purple loosestrife infestation was approximately 5.7 acres along 1.4 miles of shoreline in Capitol Lake. The established seed bank was estimated to be sufficient to support regrowth for a minimum of 10 years.

The 2001 Capitol Lake Integrated Purple Loosestrife Management Plan recommended a combination of surveillance, public education, chemical control with glyphosate spot-treatment, biological controls (insect introductions), and manual removal to eradicate the plant from Capitol Lake. This plan stressed the importance of consistency, monitoring, and data collection in effectively eradicating purple loosestrife.

From 2008 to 2015, purple loosestrife was treated with a 3 percent solution of glyphosate and plant mortality was observed within one week of application. From 2016 to 2018, a 15 percent solution of Imazapyr replaced the glyphosate treatments, and seed head removal continued. Plant mortality from imazapyr treatments was observed within one to two weeks of application (NAM 2017). Management activities in 2019 were not conducted due to an oil spill preventing lake access. In 2019, continued use of surveys, seed head removal, and imazapyr treatments was recommended for future years (NAM 2020). Survey results and observations from the management efforts are described in the following paragraphs.

From 2009 through 2012, purple loosestrife was densely populated in the South Basin and in the mitigation areas at the south end of the Middle Basin (NAM 2013). It reduced from moderately populated in 2009 to sparsely populated from 2010 through 2012 on the western side of the Middle Basin and in Percival Cove, and only up to two plants were observed in the North Basin.

In 2013, reduced densities were observed in the South Basin and the Middle Basin mitigation area. Surveys in 2013 indicated the lowest recorded populations of purple loosestrife, indicating the positive impacts of management strategies (NAM 2014). However, purple loosestrife populations increased slightly in 2014 to low densities in the South and Middle Basins, moderate densities in Percival Cove and in the mitigation areas at the southern end of the Middle Basin, and none in the North Basin (NAM 2016).

The 2015 surveys indicated a continued increase in purple loosestrife populations to high densities in the South Basin and Percival Cove, moderate in the mitigation area at the south end of the Middle Basin, and low in the Middle and North Basins (NAM 2016). These density ratings did not change in 2016, 2017, or 2018 (NAM 2017, 2018). No assessment or survey was completed in 2019 due to an oil spill that prohibited access to the lake (NAM 2020).

Yellow Flag Iris

Yellow flag iris (*Iris pseudacorus*) is a Class C Washington State Listed Noxious Weed (Washington State Noxious Weed Control Board 2020) that grows along shorelines in shallow water in wetlands, and at the edges of lakes or slow-moving rivers. The plant is perennial, grows 2 to 3 feet tall, and produces pale to dark yellow flowers with brownish purple mottled markings. Yellow flag iris spreads by rhizomes and by seed to form large clumps that outcompete native vegetation, reducing habitat complexity for wildlife. Rhizome fragments may re-root and form



Yellow Flag Iris (WNWCB 2020)

new plants, enabling the plant to further spread. The plant is toxic to livestock, not palatable to most wildlife, and its resin may irritate human skin. Dense clumps of yellow flag iris may inhibit access to shorelines for recreation. Management of yellow flag iris is costly, requiring repeated monitoring and removal efforts to prevent its spread.

Distribution and Abundance Within the Study Area

Yellow flag iris is known to have dominated wetland plant communities in Capitol Lake in 1992 and very likely existed in the area prior to this date (GA 1992). Survey results and observations from the management efforts are described in the following paragraphs.

In 2008, survey records indicate approximately 500 yellow flag iris plant clusters on Enterprise Services property in Capitol Lake. Yellow flag iris was densely populated in the South and Middle Basin and in Percival Cove, and moderately populated in the North Basin (NAM 2013). A total of 3,015 square feet of dense yellow flag iris colonies in Heritage Park, Percival Cove, and the Interpretive Park wetland ponds at the south end of the Middle Basin were treated in 2008 with glyphosate.

The number of yellow flag iris plant clusters significantly reduced to 250 in 2009 and to 200 in 2010; however, yellow flag iris remained densely populated in the South Basin and Percival Cove. The population on the eastern shoreline of the Middle Basin was significantly reduced in comparison to 2008. The density of yellow flag iris in the Middle Basin overall and in the North Basin in 2009 was moderate (NAM 2013).

In 2011, the survey team transitioned from tracking groups of plants in previous years to tracking individual plants. Approximately 450 yellow flag iris plants were observed in 2011, increasing the density in the South, Middle, and North Basin and in Percival Cove. However, plants were generally smaller and immature compared to those observed in 2008 and 2009, indicating some success in the glyphosate treatments despite the remaining presence of dense populations (NAM 2013).

In 2012 and 2013, approximately 100 yellow flag iris plants were observed at a moderate density in the South Basin, Percival Cove, and in the North Basin, and a high density in the Middle Basin. The decrease in the number of plants from 2011 to 2012 suggest the success of herbicide and manual control efforts (NAM 2013). However, the number of observed plants doubled to 200 in 2014 along with an increase to high density in the South Basin and Percival Cove (NAM 2014). The increase in yellow flag iris in 2014 may have resulted from higher water levels during the growing season, providing favorable conditions for the growth of the plant (NAM 2014).

The number of yellow flag iris plants observed on Enterprise Services property was relatively consistent at 135 plants in 2015, 142 plants in 2016,148 plants in 2017, and 125 plants in 2018 (NAM 2016, 2017, 2018). Overall, yellow flag iris was densely populated in the South and Middle Basins, particularly on the eastern shoreline of the Middle Basin, while the population density was moderate in Percival Cove and low in the North Basin. It was concluded that yellow flag iris populations could be further reduced if plants on private properties were either treated with imazapyr or manually removed.

Figure 4-3 presents the most current map of yellow flag iris distribution in 2018 with observation points colored light green (NAM 2018). As described above, yellow flag iris was documented as most abundant along the east shoreline of the Middle Basin, but was present in all three lake basins, the Interpretive Park wetlands, and Percival Cove. Yellow flag iris is not likely significantly impacting native wildlife or recreation in the Capitol Lake Basin based on its current abundance and the emergent plant habitat diversity.

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Figure 4.3 Yellow Flag Iris Distribution in Capitol Lake in 2018



Legend

—— Study Area 🛛 🗕 Yellow Flag Iris

Management Approaches

From 2008 to 2015, yellow flag iris was surveyed annually and treated with a 3 percent solution of the aquatic herbicide glyphosate. Plant die-off was observed 1 to 2 weeks after application (NAM 2013). In 2016, treatment was transitioned to a 1.5 percent solution of imazapyr (NAM 2016). Plant die-off was observed 1 to 2 weeks after application. The imazapyr chemical treatment continued in 2017 and 2018 (NAM 2018). During each year of management, yellow flag iris was treated and counted on all Enterprise Services property adjacent to the shoreline of Capitol Lake. Seeds were removed from plants on residential properties adjacent to the shoreline of Capitol Lake, but these plants were not included in survey numbers.

Reed Canarygrass

Reed canarygrass (*Phalaris arundinacea*) is a Class C Washington State Listed Noxious Weed (Washington State Noxious Weed Control Board 2020) that grows along the margins of lakes and streams and in wetlands in wet open areas. The plant is a tall, coarse, perennial grass (Ecology 2001). Stems may grow to be 2 meters tall with leaves that are blue-green when new and strawcolored when dry, and pinkish flowers at full bloom. Reed canarygrass spreads by rhizomes and by seed to form dense monoculture colonies that outcompete native vegetation, reduce



Reed Canarygrass (WNWCB 2020)

habitat complexity for wildlife, and inhibit recreational activities along shorelines. It grows too densely to be used by waterfowl for cover and is not eaten by many wildlife species. It may provide amphibian habitat. It also slows and filters surface water during storm events, thus providing important hydrologic and water quality functions.

Distribution and Abundance Within the Study Area

Reed canarygrass is known to have dominated wetland plant communities in Capitol Lake in 1992 and very likely existed prior to this date (GA 1992). It was observed in Capitol Lake in 2004 as dominant in nearly monospecific (one species only) patches, but a map of its distribution is not available (Ecology 2020a). Reed canarygrass is not likely significantly impacting native wildlife based on the current emergent plant diversity in in the Capitol Lake Basin.

Management Approaches

Reed canarygrass has not been the focus of any documented large-scale management or monitoring efforts reviewed for this study. As a result, detailed descriptions of its presence in Capitol Lake are unavailable; however, it is very likely present in wetlands adjacent to all basins due to the lack of control efforts.

4.1.1.2 Floating-Leaved Plants

Fragrant Water Lily

Fragrant water lily (*Nymphaea odorata*) is a Class C Washington State Listed Noxious Weed (Washington State Noxious Weed Control Board 2020) that grows on the edges of lakes and ponds. The plant produces fragrant white or pinkish flowers that float on the surface of the water (Ecology 2001). Fragrant water lily has large, round floating leaves and is rooted in the sediment below the water surface. The plant spreads by seed as well as with long, thick rhizomes that grow just beneath the sediment surface and produce new plants and thread-like roots. Fragrant water lily may



produce new plants and thread-like roots. Fragrant water lily may *Fragrant Water Lily (WNWCB 2020)* form large, dense stands in water, outcompeting native aquatic vegetation and inhibiting recreational access along shorelines.

Distribution and Abundance Within the Study Area

Records of fragrant water lily in Capitol Lake do not indicate significant issues with large infestations of this plant. The first records of management efforts begin in 2008, when the GA contracted with NAM to control aquatic invasive plants in Capitol Lake. A 2008 survey found no clusters of fragrant water lily in the South Basin, 2 clusters in the Middle Basin, 15 clusters in Percival Cove, and 4 clusters in the North Basin (NAM 2013). Fragrant water lily very likely existed prior to this 2008 survey, although no records confirming its prior presence were available for this study.

Figure 4-4 presents the most current map of fragrant water lily distribution in 2018 with observation points colored light blue (NAM 2018). Fragrant water lily plants were observed at three locations in the North Basin (north, southwest, and southeast nearshore areas) and multiple locations in Percival Cove, but not in the Middle or South Basins. Fragrant water lily is not likely significantly impacting native wildlife or recreation in the Capitol Lake Basin based on its current abundance and the aquatic plant habitat diversity.

Management Approaches

Management of fragrant water lily is costly, requiring repeated monitoring and removal efforts to prevent its spread. From 2008 to 2018, fragrant water lily was routinely surveyed and managed by NAM with repeated cuttings of leaves and tops of stems to stress the plants. During each year of management, NAM repeatedly cut and counted all clusters of fragrant water lily on Capitol Lake (NAM 2013, 2018).

No clusters were found in the South Basin or the Middle Basin by any survey from 2010 through 2018. Slightly fewer clusters were observed in Percival Cove, ranging from 11 in 2010 to a minimum of 6 in 2018. Clusters in the North Basin ranged from one to five until a maximum of seven were observed in 2018.

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Figure 4.4 Fragrant Water Lily Distribution in Capitol Lake in 2018



Legend

— Study Area 🔹 🗕 Fragrant Water Lily

4.1.1.3 Submersed Plants

Eurasian Watermilfoil

Eurasian watermilfoil (*Myriophyllum spicatum*) is a Class B Washington State Listed Noxious Weed (Washington State Noxious Weed Control Board 2020) and a High Priority invasive species (WISC 2020c) that grows submersed below water surfaces. The plant has feather-like underwater leaves, emergent flower spikes, and many fibrous roots (Ecology 2001). Roots may form on broken plant fragments, enabling the plant to spread by plant fragments in addition to spreading by rhizomes. The abundance of viable plant fragments allows this



plant to rapidly spread and colonize new areas and it commonly *Eurasian Watermilfoil (WNWCB 2020)* forms dense, thick mats early in the growing season due to its rapid growth rate of up to 1 foot per week. These mats reduce sunlight and oxygen in underlying waters, thus degrading water quality, outcompeting native vegetation, decreasing habitat quality for native fish species, and inhibiting recreational activities. Management of Eurasian watermilfoil is costly, requiring repeated monitoring and removal efforts to prevent its spread.

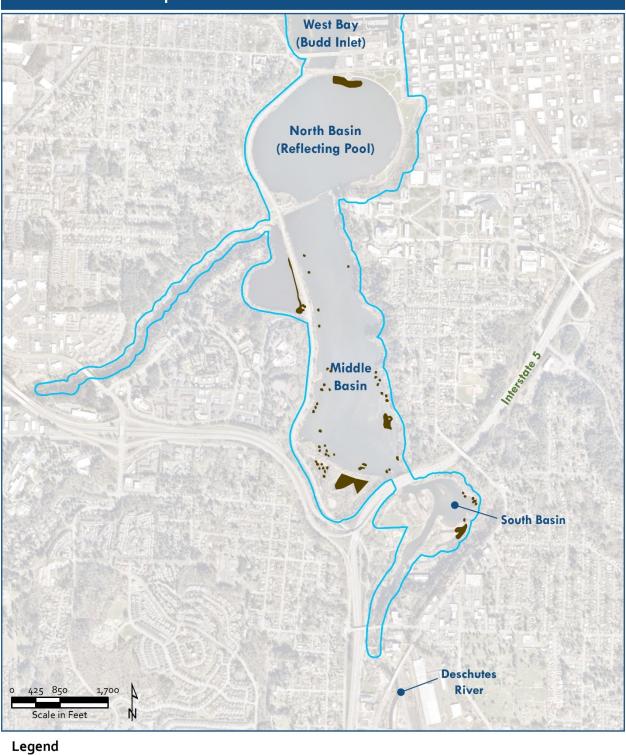
Distribution and Abundance Within the Study Area

A representative of the Thurston County Noxious Weed Board first reported Eurasian watermilfoil in Capitol Lake in September 2001 (GA 2004). The Thurston County Department of Water and Waste Management subsequently conducted an aquatic plant survey of the lake in October 2001 and identified Eurasian watermilfoil in the Middle and North Basins of the lake (Aquatechnex 2003).

Figure 4-5 presents the most current map of Eurasian watermilfoil distribution in 2018 with observation points and areas colored red (NAM 2018). Individual plants and patches of plants were observed in all three basins, the Interpretive Park wetland ponds, and Percival Cove. The large patch observed in the North Basin adjacent to the dam had not been observed in previous surveys. Although Eurasian watermilfoil is a high-priority species based on its aggressive growth and potential impact on native species, it not likely significantly impacting native wildlife or recreation in the Capitol Lake Basin based on its current abundance and the aquatic plant habitat diversity. Native submersed plants currently impact boat navigation of the lake in summer because they grow to water surface over most of the lake area, as noted during lake water quality monitoring for the EIS project (N. Maas, Herrera Environmental Consultants, Inc., personal communication).

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Figure 4.5 Eurasian Watermilfoil Distribution in Capitol Lake in 2018



Study Area

Eurasian Watermilfoil

Management Approaches

In response to the discovery and survey, the Capitol Lake Adaptive Management Plan (CLAMP) steering committee adopted an Invasive Aquatic Vegetation Management Plan in 2002 that included a triclopyr monitoring plan with preapplication testing in Capitol Lake and Budd Inlet to establish baseline data so that the effects of treatment activities could be fully understood as maintenance efforts continue.

On July 19, 2004, Aquatechnex applied triclopyr to Eurasian watermilfoil in the South Basin, Middle Basin, Percival Cove, and in the Interpretive Park wetland ponds. On July 29, 2004, Aquatechnex applied triclopyr to Eurasian watermilfoil in the North Basin. The Washington Department of Agriculture monitored the application process and noted that triclopyr was effective in killing Eurasian watermilfoil, dissipated quickly, and did not harm native aquatic vegetation. Eurasian watermilfoil plants in the South Basin were not completely eradicated due to flow from the Deschutes River limiting the contact time of triclopyr on treated vegetation in this area. Remaining plants were hand pulled in August of 2004 (GA 2004).

In the spring and summer of 2005, the GA monitored Capitol Lake to identify any Eurasian watermilfoil plants. Fewer than 100 plants were identified in the South Basin and in the Interpretive Park wetland ponds. Divers hand pulled these plants and used a water vacuum to capture any floating fragments. Monitoring was repeated in 2005 after manual removal, and no regrowth was observed.

From 2007 to 2018, NAM annually surveyed and manually removed Eurasian watermilfoil by boat and/or snorkel team (NAM 2018). In 2007, 1,386 plants were removed from the South and Middle Basins. The number of plants removed from these basins decreased annually thereafter, to 171 plants in 2008, 132 plants in 2009, 51 plants in 2010, 36 plants in 2011, 7 plants in 2012, and 6 plants in 2013 and 2014. The number of plants removed then increased to 56 plants in 2015 and 2016, 80 plants in 2017, and 105 plants in 2018.

Eurasian watermilfoil was first reported in Percival Cove as one plant in 2009 and continued to increase in abundance until hand removed by divers in 2012 and 2013 when 3,500 and 1,320 pounds were removed, respectively (ACERA 2013). Fewer than 20 plants were removed each year from Percival Cove in 2015 through 2018 (NAM 2018).

Eurasian watermilfoil was first reported in the west Interpretive Park wetland pond as 27 plants in 2009 that decreased to 8 plants in 2015, and ultimately increased to 15 to 20 plants in 2018 (NAM 2018). Dense plants were also observed in the east Interpretive Park wetland pond in 2015.

Eurasian watermilfoil was first reported in the North Basin as one plant in 2010, but not again until 2018 when a patch was observed just east of the 5th Avenue Dam (NAM 2018).

In 2019, an oil spill prevented access to Capitol Lake for vegetation surveys and management. NAM reported a high likelihood that Eurasian watermilfoil would continue to expand, consistent with monitoring results in 2018. NAM recommended that maintenance strategies for 2020 include in-depth surveys, individual plant marking, diver dredging, and shading cloths over individual plants (NAM 2020).

The GA implemented a Eurasian watermilfoil weevil (*Euhrchipsis lecontei*) biocontrol strategy for the east Interpretive Park wetland pond in August 2009. These efforts were monitored in 2009 and 2010, and an additional control effort was implemented by installing bottom barriers. Initial monitoring results at the time indicated that weevils could survive stocking, overwinter, and may grow to populations that control Eurasian watermilfoil in small water bodies. Results were somewhat inconclusive due to site disturbance (GA 2010). Weevils survived the saltwater flush of Capitol Lake in 2010. Over time, the Eurasian watermilfoil population reestablished in the east pond (NAM 2018), indicating a lack of long-term success for this control strategy.

Bottom barriers were effective in eliminating Eurasian watermilfoil growth in the Interpretive Park wetland ponds and were first installed in the main lake body in 2009 and in the west pond for clusters of multiple plants. A total of 675 square feet of bottom barriers were installed in Capitol Lake in 2009 and were observed to have remained intact and be effective in 2011. In 2010, an additional 225 square feet of bottom barrier was installed in the west pond (NAM 2011). Additional information on bottom barrier removal and maintenance was not available for this study; however, the increase in Eurasian watermilfoil in recent years indicates a lack of ongoing maintenance and reduced effectiveness of the bottom barriers. Without maintenance or replacement, the effectiveness of bottom barriers is reduced by displacement, degradation, or sediment accumulation.

Curlyleaf Pondweed and Ecological Effects

Curlyleaf pondweed (*Potamogeton crispus*) is a Class C Washington State Listed Noxious Weed (Washington State Noxious Weed Control Board 2020) that grows submersed below water surfaces in shallow to deep and still to flowing water. Leaves are olive-green to reddish-brown, oblong, stiff, and translucent with wavy edges and three prominent veins (Ecology 2001). Curlyleaf pondweed may grow stems up to 90 centimeters long and small flowers with four petal-like leaves that emerge above the water surface. The plant spreads by rhizomes and by seed. Curlyleaf pondweed may form dense surface mats that



Curlyleaf Pondweed (WNWCB 2020)

reduce sunlight and oxygen in underlying waters, thus degrading water quality, outcompeting native vegetation, decreasing habitat quality for native fish species, and inhibiting recreational activities.

Distribution and Abundance Within the Study Area

Curlyleaf pondweed was observed and documented in Capitol Lake in 2001, 2003, 2004, and 2006 (Ecology 2020a), and likely existed prior to these records. Its density was not noted for the 2006 survey, but results of the 2006 survey indicate its presence as sparse with a wide patchy distribution that was most abundant in the south end of the lake. Curlyleaf pondweed is not likely significantly impacting native wildlife or recreation in the Capitol Lake Basin based on its current abundance and the aquatic plant habitat diversity.

Management Approaches

Management of curlyleaf pondweed can be costly; however, it is only occasionally problematic in Washington. Due to its low abundance this plant is not considered to be a problematic species in Capitol Lake, and as such, it has not been the focus of any management efforts.

4.1.2 Presence in Upstream Waters

Purple loosestrife has been observed and documented upstream of Capitol Lake on the Deschutes River (Thurston County 2020b). Although no other aquatic invasive plants have been mapped on the Deschutes River or Percival Creek, it is possible that other aquatic invasive plant species are present upstream of the lake. No aquatic invasive plant monitoring or management efforts are known to have occurred on the Deschutes River or Percival Creek.

4.1.3 Presence in Budd Inlet

Purple loosestrife has been documented downstream of Capitol Lake in Budd Inlet (Thurston County 2020b). No other aquatic invasive plant species are mapped in Budd Inlet. With the exception of several species of cordgrass (*Spartina* spp.) no other Class A noxious weeds are salt tolerant; therefore, they are unlikely to be found in Budd Inlet.

4.2 AQUATIC INVASIVE ANIMALS

Aquatic invasive animals found in Capitol Lake and the surrounding water bodies include four invertebrate, five fish, one waterfowl, and one mammal species. Table 4-2 describes their classification on the WDFW AIS database (WDFW 2020b) and distribution and relative abundance in the study area.

Scientific/ Common Name	Classification ^a	Water Body	Relative Abundance	Source
Invertebrates				
Potamopyrgus antipodarum New Zealand mudsnail	Prohibited High Priority	Capitol Lake	20,000 snails per square meter in limited areas of the North Basin	Johannes 2011
<i>Corbicula fluminea</i> Asiatic clam	Not listed	Capitol Lake	Approximately 287 snails	Herrera 2004
<i>Radix auricularia</i> European ear snail	Not listed	Capitol Lake	Approximately 38 snails	Herrera 2004

Table 4-2. Aquatic Invasive Animals Observed in the Study Area.

CAPITOL LAKE — DESCHUTES ESTUARY

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Scientific/ Common Name	Classification ^a	Water Body	Relative Abundance	Source
Fish				
<i>Cyprinus carpio</i> Common carp	Regulated	Capitol Lake	Fewer than 200 fish	Herrera 2004
<i>Ameiurus nebulosus</i> Brown bullhead	Not listed	Capitol Lake	Fewer than 50 fish	Herrera 2004
<i>Micropterus salmoides</i> Largemouth bass	Not listed	Capitol Lake	Fewer than 200 fish	Herrera 2004
<i>Perca flavescens</i> Yellow perch	Not listed	Capitol Lake	Fewer than 50 fish	Herrera 2004
Waterfowl				
<i>Branta canadensis</i> Canada goose	Not listed	Capitol Lake	142 birds	USDA 2018
Mammals	·			
<i>Myocastor coypus</i> Nutria	Prohibited High Priority	Capitol Lake	Fewer than 25 animals	Herrera 2004

Classification by the Washington State Department of Fish and Wildlife Aquatic Invasive Species Database (WDFW 2020b) or High Priority invasive species status (WISC 2020c).

4.2.1 Presence in Capitol Lake

The biology, documented presence in Capitol Lake, potential economic and ecological impact of the species, and possible management actions for invertebrates, amphibian, fish, waterfowl and mammal are described below.

4.2.1.1 Invertebrates

New Zealand Mudsnail

New Zealand mudsnail (*Potamopyrgus antipodarum*) is a Prohibited (WDFW 2020b) and High Priority (WISC 2020c) invertebrate AIS. It is a very small freshwater snail with an elongated shell, approximately 4 to 6 millimeters (mm) with five to eight right-turning whorls and an operculum that can seal the opening of the shell. The operculum allows the mudsnail to protect itself from short-term exposure to most chemicals, from extended periods of desiccation, and allows them to survive outside water for long periods of time. New Zealand mudsnails are grazers and feed on attached periphyton and decaying plant



New Zealand mudsnail (WISC 2020a).

and animal material (Haynes and Taylor 1984). They are often found in densities up to 500,000 snails per square meter (Richards 2004; Hall et al. 2003).

The New Zealand mudsnail is dioecious (separate male and female sexes) and bears live young (Wallace 1978). Female mudsnails can reproduce sexually or asexually, producing cloned genetically identical offspring. Thus, one female is enough to initiate a new population, and all introduced populations outside of New Zealand are entirely clonal (Zaranko et al. 1997). New Zealand mudsnails are found in shallow freshwater and brackish water ecosystems. In general, they have a broad environmental tolerance to diverse ranges of temperature, osmotic concentrations, flows, substrates, and disturbance regimes, although clonal lineages may have varied tolerances to environmental conditions (Dybdahl and Lively 1995; Jokela et al. 1999; LeClair and Cheng 2011). Due to their ability to survive outside the aquatic environment for several weeks to months, new populations can be established through transportation on contaminated boots, gear, and equipment.

While the New Zealand mudsnail has no natural predators in Washington, laboratory studies show that rainbow trout and steelhead will feed on New Zealand mudsnails; however, fish derive little or no energy value from eating snails because the snails are capable of passing through the fish's digestive system alive and intact (Bruce et al. 2009). Although there have been documented incidences of New Zealand mudsnails found in Chinook salmon stomach contents, they are not common or important prey items (Bersine et al. 2008).

Many of the ecological impacts of the New Zealand mudsnail can be attributed to the extreme rate of population growth and high densities it can reach (Alonso and Castro-Díez 2008). The New Zealand mudsnail consumes high amounts of periphyton (i.e., algae attached to rocks and other substrates) and excretes high amounts of total ammonia, dominating both the carbon and nitrogen cycles (Hall et al. 2003; Hall et al. 2006; Alonso and Castro-Díez 2008). In addition to outcompeting native species for natural resources, their ability to withstand highly variable environmental conditions allows New Zealand mudsnails to take advantage of changing environmental conditions, including climate change, to further spread and outcompete native species (Alonso and Castro-Díez 2008). By outcompeting native species, the New Zealand mudsnail reduces prey species for native fish, resulting in reduced body weight and health of native salmonids (Richards 2004; Vinson and Baker 2008; Alonso and Castro-Díez 2012). The United States Fish and Wildlife Service (USFWS) concluded that after 20 years since its invasion of the Pacific Northwest, there was limited evidence that the New Zealand mudsnail had ecosystem level impacts, there had been no cascading ecological effects, and there had been multiple unexplained population crashes (Draheim 2015).

Biofouling is the major economic impact associated with the introduction of New Zealand mudsnails, in general. New Zealand mudsnails have been documented to pass through water pipes and emerge from domestic taps (Ponder 1988) and can block water pipes and meters (Zaranko et al. 1997). Other economic costs are related to the vulnerability of threatened or endangered species, research and development expenses incurred by agency and university personnel to prevent further spread of New Zealand mudsnails, extra steps and equipment required to prevent the further spread of invasive

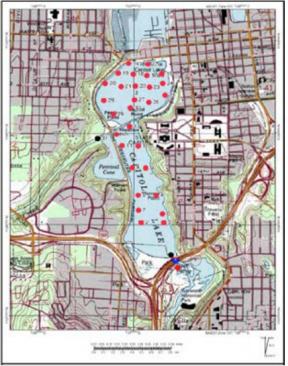
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species, and contamination of private hatcheries and subsequent regulation and prohibition of their operations (Proctor et al. 2006).

Distribution and Abundance Within the Study Area

The invasion of the New Zealand mudsnails in fresh and brackish water habitats spans across Europe, Australia, Japan, and North America (Johannes 2013). In the United States, an all-female population of New Zealand mudsnails was discovered in the middle Snake River, Idaho, in 1987 (Taylor 1987). Since its discovery, they have spread to 10 western states, including Washington in 2002 (Davidson et al. 2008).

New Zealand mudsnails were first observed in Capitol Lake in 2009 (Bartleson 2010; Bensen 2010; Johannes 2010). A 2011 survey of Capitol Lake detected New Zealand mudsnails in 30 of the 31 sites investigated (Johannes 2011). A 2013 survey of 83 sites around Capitol Lake and the surrounding area, detected New Zealand mudsnails in three locations: North Basin, Middle Basin, and the Deschutes River, but did not extend into nearby creeks and tributaries (Johannes 2013; USGS 2020). No New Zealand mudsnails were observed in the five new sites in streams and lakes that were surveyed within a 5-mile radius of Capitol Lake in 2015, and populations detected in the 2013 survey were



New Zealand mudsnail distribution in Capitol Lake in 2013 (red points positive, black points negative) (Pleus 2016)

still observed in low numbers (Johannes 2015). No New Zealand mudsnails were detected at the mouth of Percival Creek (Johannes 2015), suggesting that they had not migrated further upstream into Percival Creek in the 6 years from when they were originally observed in Capitol Lake to the most recent mudsnail survey. No mudsnail sightings in Percival Creek or other nearby waters have been reported since the 2015 survey (WISC 2020a), suggesting that the spread of mudsnails from the lake has been very limited over the past 10 years.

New Zealand mudsnails are currently affecting recreational opportunities in the project area, but their impact on native wildlife is unclear. Several native species of snails are also abundant in Capitol Lake, and the New Zealand mudsnail population has not overtaken the benthic community in the lake as was expected (Pleus 2016).

Management Approaches

Management of New Zealand mudsnails occurs through natural environmental conditions and physical and chemical treatments. Given the species persistence, prevention of repeated introductions and further spread relies on public outreach and education. Literature review, preliminary testing, and

professional judgement were used to identify a set of factors that may affect the spread and mortality of New Zealand mudsnails in Capitol Lake. These include:

- Freezing
- Heat and desiccation
- Saltwater backflush
- Depth and water temperature
- Presence of structures
- Substrate grain size
- Calcium concentration and water hardness
- Introduced chemical agents

Freezing can increase the mortality response when the lakebed is drained and exposed to hard freezing weather conditions for a few consecutive days. The use of this measure is highly dependent upon weather conditions that must be both cold and dry without insulating snow, which is an unusual combination of conditions for the Capitol Lake area. The effectiveness of freezing is also limited by short periods of freezing and the resulting shallow freezing depths in Puget Sound lowland lakes. Enterprise Services lowered the level of the lake between December 15 and 18, 2016 (Enterprise Services 2016). New Zealand mudsnail mortality varied depending on location. For example, mortality near Marathon Park was approximately 50 percent, whereas areas along Powerhouse Road were approximately 90 percent. These differences were attributed the proportion of the survey areas exposed to freezing conditions where less mortality was observed where more area was below ice cover (Enterprise Services 2016).

Heat and desiccation through local weather conditions is more frequently achieved than dry and freezing conditions. However, this seasonally dependent action requires several consecutive hot-dry days and has been shown in productive lakes to cause nuisance odors from decaying algae and aquatic plants and animals. Both the freezing and heat factors are limited in Capitol Lake by the mild climate and constant inflow from the Deschutes River.

In an experiment that tested New Zealand mudsnail tolerance to osmotic and thermal shock, researchers found the snails had significantly higher mortality at higher salinities that increased with higher temperatures (Paolucci and Thuesen 2020). While increasing the salinity may increase mortality, an experimental backflush to treat New Zealand mudsnail populations in Capitol Lake in 2010 found the treatment reduced the entire macroinvertebrate community (Adams 2010). Researchers speculated that after the lake returned to normal freshwater conditions, the backflush could result in a larger mudsnail population as a product of rapid reproduction and newly available resources that are no longer consumed by competing species (Adams 2010).

New Zealand mudsnails have been shown to tolerate salinities greater than the 26 to 30 parts per thousand (ppt or practical salinity unit) observed in Budd Inlet. In addition, salinity tolerance has been

shown to increase with marine water exposure. New Zealand mudsnails in the Columbia River estuary are more tolerant of acute salinity stress with LC50 values (lethal concentration causing 50 percent mortality) averaging 38 ppt salinity versus only 22 ppt salinity for mudsnails from a freshwater source (Devils Lake) (Hoy et al. 2012).

How water depth and temperature affect the New Zealand mudsnail distribution in Capitol Lake has been studied by WDFW but is not well understood. Given the resiliency of this species, it is unlikely that manipulation of these factors by regulating lake outflow would significantly impact the mudsnail population.

The presence of structures in the water column may influence the success of New Zealand mudsnail populations, but this factor is not well understood. Anecdotal information from Capitol Lake suggests the snails prefer locations with legacy piers, debris, and other uneven substrate features. Structures are presumed to provide shelter from various threats and an ability to regulate ambient conditions such as temperature or light.

Substrate grains size is also poorly understood regarding population success. New Zealand mudsnail have been found to thrive in silt and gravel dominated environments. Capitol Lake's Middle and North Basins are almost entirely a fine silt substrate. The southern end of the system is dominated by gravel. Presence/absence surveys have not found New Zealand mudsnail in the southern, graveled portion of the lake, but it is not clear if substrate grain size is the factor limiting New Zealand mudsnail migration to the southern portion of the lake.

Calcium concentration and water hardness are linked with New Zealand mudsnail success. Literature review has identified a pattern of greater snail population success in waters with a high mineral content. Most water systems influenced by urban development have a greater hardness than their natural counterparts. As with other factors discussed here, it is difficult to conceive how calcium concentration and water hardness could be altered to successfully affect the existing snail infestation.

Efficacy information has been identified for various chemical agents through literature review and recent laboratory analyses. Areas where chemical eradication may be possible include water bodies that can be isolated from the drainage, such as small lakes and ponds, irrigation canals, and fish hatcheries. Two chemical agents examined for use in Capitol Lake are Bayluscide (with niclosamide as the active ingredient) and sodium chloride. Bayluscide acts quickly, killing the New Zealand mudsnail before they have a chance to respond or find protection (Nautilus Environmental 2011). The lethal action of sodium chloride is much slower, allowing the snail to close its operculum and wait for the toxic level of the introduced agent to dissipate. Additional lab analysis for sodium chloride treatments would be needed to better understand how it might perform in a field application. Neither chemical is currently allowed for aquatic use under the Aquatic Invasive Species Management Permit, but application of either chemical may be allowed by an experimental use permit or addition of the chemical to the existing permit as part of its 5-year update, which is due in 2021, 2026, etc. (Ecology 2020b).

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In addition to treatment efforts to control New Zealand mudsnail population growth and spread, public outreach and education can help to prevent the spread by human activity. Signage warns recreational users at Marathon Park of New Zealand mudsnail infestations. Although educational outreach is a helpful first step in encouraging additional public monitoring and control, signs alone are not effective to preventing the spread of invasive species.

European Ear Snail

The European ear snail (*Radix auricularia*), also known as the bigear radix, is a species of medium-sized freshwater snail. The shell is thin, roundly ovate, and very inflated, such that the last whorl comprises 90 percent of its volume. It is native to Europe and temperate Asia, found in slow-moving rivers and streams, ponds, lakes, and wetlands. The European ear snail has a wide tolerance for temperatures and specific conductivities (Vinarski and Serbina 2012; von Oheimb et al. 2016).

There is limited information available on the European ear snail effects on native ecosystems; however, it is likely similar to the information described above for the New Zealand mudsnail. The invasive European ear snail reduces native invertebrate diversity



European ear snail (Kipp et al. 2019).

by consuming large amounts of primary producer biomass and altering ecosystem functions (Hall et al. 2003; Riley et al. 2008). In some environments, European ear snail may prey on juvenile snails, directly impacting native snail populations (Deng 1997).

Distribution and Abundance Within the Study Area

Originally, the European ear snail was likely introduced accidentally on plants imported to North American greenhouses (Kipp et al. 2020). Subsequent introductions may have occurred through releases from aquaria and the European ear snail is now found in 22 of the 50 United States, including California, Idaho, Montana, Oregon, and Washington (Kipp et al. 2019). The European ear snail was first documented in Capitol Lake in 2003 (Herrera 2004) but was not found in any sites investigated in or around Capitol Lake during the 2011 or 2013 surveys (Johannes 2011; Johannes 2013). European ear snail is not likely significantly impacting native wildlife in the Capitol Lake Basin based on its current abundance and the aquatic habitat diversity.

Management Approaches

Although there is little information on species-specific control and management efforts for European ear snail, preventing the spread by educating the public and applying chemical treatments are the most likely options to reduce their impacts.

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Asiatic Clam

The invasive Asiatic clam (*Corbicula fluminea*) is a freshwater bivalve that originates from Asia (Britton and Morton 1979). The species is opportunistic, adapting to a wide range of freshwater ecosystems and utilizes both filter- and deposit-feeding strategies, feeding on phytoplankton and zooplankton (Way et al. 1990). The Asiatic clam is 2 to 3 cm in size and has an ovaltriangular shell that ranges from dark olive green to black (Foster et al. 2019). Reproductive strategies of the Asiatic clam include sexual reproduction with both sexes or hermaphrodites and several other unusual reproductive features, ranging from



Asiatic clam (Foster et al. 2019).

oviparity and ovoviviparity to euviviparity (Korniushin and Glaubrecht 2003; CABI 2020a).

Oxygen availability may be one of the most critical factors affecting their population growth and distribution, with mature individuals requiring greater than 70 percent dissolved oxygen saturation for survival (McMahon 1979). Due to this high oxygen demand, Asiatic clams typically inhabit well-oxygenated streams and lake shallows. The maximum tolerable salinity for Asiatic clams is approximately 13 parts per thousand, but this concentration can only be withstood for a short time (Aguirre and Poss 1999). Although they can survive in water temperatures as low as 2° Celsius, the clams prefer warmer habitats (McMahon 2002).

Ecological impacts include altering the physical, chemical, and biological characteristics of the invaded water body. Their feeding versatility and rapid rate of reproduction contribute to the clam's ability to quickly establish colonizing populations in a broad range of aquatic habitats and is capable of shifting zooplankton communities in favor of copepods by selectively removing rotifers (Beaver et al. 1991), altering the biota of a habitat and leading to the disruption of benthic and pelagic community structure (Sickel 1986). This can affect aquatic birds and increase the frequency of nuisance species and algal blooms. Drastic reductions in native species populations are also observed due to direct competition with Asiatic clams for limited resources (Devick 1991).

Distribution and Abundance Within the Study Area

Since the first known introduction in 1937 to the Columbia River, the Asiatic clam has been observed in 46 states, the District of Columbia, and Puerto Rico, and has spread to 45 hydrologic units in Washington State (USGS 2020). The mechanism for dispersal within North American is unknown (Foster et al. 2019). In 2003, approximately 287 Asiatic clams were documented in the North Basin along the west shoreline of Capitol Lake at a depth of 7 feet (Herrera 2004). Despite its known presence in Washington, it is only a moderately successful invader and does not occur in great densities (Johannes 2010, 2011, 2013). The Asiatic clam was found at one location in Black Lake during the 2011 survey of sites outside Capitol Lake (Johannes 2011) and at three sites as either shell fragments or complete dead shells in Black Lake during the 2013 survey of sites outside Capitol Lake to during the 2013 survey of sites outside Capitol Lake into

Capitol Lake (Johannes 2013; USGS 2020). The Asiatic clam is not likely significantly impacting native wildlife in the Capitol Lake Basin based on its current abundance and the aquatic habitat diversity.

Management Approaches

Few options exist for the management of established Asiatic clam populations due to the persistence of larval stages in the water column. Chemical treatments (chlorination and bromination) are used for both juveniles and adults, but these treatments can cause severe environmental damage in open water systems if administered incorrectly (INDNR 2009). Physical and benthic barriers and diver-assisted suction could be used for species control (Wittmann et al. 2012; State of Michigan 2020), as well as dry ice applications (Coughlan et al. 2018), and open-flame burn treatments of exposed Asiatic clam beds (Coughlan et al. 2019). Dredging is also a treatment option, but also removes native organisms and can miss veligers or juveniles (State of Michigan 2020).

4.2.1.2 Fish

Common Carp

The common or European carp (*Cyprinus carpio*) can be identified by several features, including large scales, two barbels on each side of the mouth, and the first dorsal and anal fin spines are serrated. Most carp are bronze-gold to golden yellow on the sides and yellowish white on the belly. This species generally inhabits lakes, ponds, and the lower sections of rivers with moderately flowing or standing water, but is also



Common carp (Nico et al. 2020).

known from brackish-water estuaries, backwaters, and bays. In its native range, the species occurs in coastal areas and estuaries withstanding salinities up to 14 ppt. In the United States, the common carp has been observed in waters with salinities as high as 17.6 ppt (Nico et al. 2020), which is much less than the 26 to 30 ppt range observed in Budd Inlet.

Larval common carp feed primarily on zooplankton. Feeding habits of juvenile and adults include benthic organisms, vegetation, detritus, plankton, chironomids, small crustaceans, and gastropods (Nico et al. 2020). The common carp are an important seed dispersal vector for aquatic plants.

The feeding habits and movements of the common carp are quite disruptive, often disturbing sediments and increasing turbidity. These actions can retard the growth of submerged aquatic vegetation, disturb spawning and nursery areas of native fish, and affect food sources for native waterfowl (Nico et al. 2020). In addition to competing with native fishes for natural resources, common carp also prey on the eggs of other fish species and may be responsible for the decline of native species (Nico et al. 2020).

Distribution and Abundance Within the Study Area

In Washington, the common carp are found in 23 watersheds, including the Deschutes watershed and greater Puget Sound area (USGS 2020). Common carp were observed in Capitol Lake during a 2003

survey but were not identified as a dominant species (Herrera 2004). Common carp is not likely significantly impacting water quality or native wildlife in the Capitol Lake Basin based on its current abundance.

Management Approaches

Due to impacts of the common carp, concentrated efforts in other areas to permanently eliminate fish include trapping, seining, and poisoning, but few were successful (NPS 2019). In addition, common carp abundance has been successfully reduced using electrofishing in Seattle's Green Lake; and lake associations in Thurston County have held carp fishing derbies to remove carp (R. Zisette, Herrera Environmental Consultants, Inc., personal communication).

Brown Bullhead

The brown bullhead (*Ameriurus nebulosus*) is a member of the catfish family and is identified by the presences of strong barbs or serrations on the back edge of its pectoral spines, and pigmentation in the chin barbels. They are often found in muddier, warmer waters and can tolerate high water temperatures and low dissolved oxygen levels that would be lethal to most other fish.

Although full estimates or documentation of the impact from brown bullhead on habitat and native populations are lacking, the brown bullhead may increase the physical disturbance in freshwaters habitat during their benthic feeding activities. The aggressive

foraging may be necessary to dislodge certain benthic prey items, which in-turn can increase turbidity and lead to altered productivity and nutrient cycling. In addition to outcompeting native species for natural resources, the brown bullhead preys on small fishes and invertebrates, reducing native populations.

Distribution and Abundance Within the Study Area

Brown bullhead are native to the Atlantic and Gulf Slope drainages from Nova Scotia and New Brunswick to Mobile Bay, Alabama, and east to the Apalachicola and Mississippi Rivers (Fuller and Neilson 2020a). Introduced brown bullhead may have been intentionally stocked around the United States for food and sport and has been reported in Washington as early as 1880 (Fuller and Neilson 2020a). After being introduced, the brown bullhead become established. During a study to inventory invertebrate and vertebrate populations in Capitol Lake, fewer than 50 brown bullheads were observed (Herrera 2004). Brown bullhead is not likely significantly impacting native wildlife in the Capitol Lake Basin based on its current abundance.

Management Approaches

Previous efforts to control brown bullhead populations include enhancing native predation. Brown bullheads are the most susceptible to predators in the developmental stages. Other methods of



Brown bullhead (Fuller and Neilson 2020a).

eradication using physical removal and chemical agents; however, both those treatment options affect native species, as well.

Largemouth Bass

The largemouth bass (*Micropterus salmoides*) occupy a variety of habitats ranging from large lakes, rivers, and reservoirs, to smaller water bodies including ponds and creek pools. Largemouth bass are opportunistic feeders that exploit a variety of prey and are highly piscivorous as adults. They differ from the smallmouth bass with a darker greenish-black color, prominent dark horizontal band along the midline, and the large maxillary that extends fully past their eye.



Largemouth bass (Fuller et al. 2020b).

Largemouth bass impact populations of small native fish directly through predation, sometimes resulting in the decline or extirpation of native species (Fuller et al. 2020b). The increased predation at one or more lower trophic levels could result in altered ecosystem processes (e.g., primary productivity and nutrient cycling) (CABI 2020b).

Distribution and Abundance Within the Study Area

Largemouth bass have been intentionally stocked for the sports fishery in several states and countries. In Washington, the largemouth bass has been intentionally stocked since the 1890s and is currently found in 58 watersheds including the Nisqually River watershed (Fuller et al. 2020b). The inventory of invertebrates and vertebrates in Capitol Lake observed fewer than 200 largemouth bass (Herrera 2004). Largemouth bass are also present in several small lakes surrounding the watershed (USGS 2020). Largemouth bass is not likely significantly impacting native wildlife in the Capitol Lake Basin based on its current abundance.

Management Approaches

Management options for largemouth bass populations focus on harvest regulations. When largemouth bass become a nuisance species, more liberal fishing regulations help control population to reduce predation on native fish species.

Yellow Perch

Yellow perch (*Perca flavescens*) are identified by their golden-yellow coloring, although the intensity of the color may vary with age and water clarity. This species has a high tolerance for low oxygen levels and acidification and is known to survive winterkill (CABI 2020c). Yellow perch prefer clear waters and remain close to shore near vegetation (CABI 2020c).



Yellow perch (Fuller and Neilson 2020b).

Yellow perch compete with native salmonid species for food and likely prey on native juvenile fish species. Nonnative predators, including yellow perch, reduce the abundance and diversity of native prey species in several Pacific Northwest rivers (Fuller and Neilson 2020b). Predation of yellow perch on small native species may also alter the ecosystem, changing the structure of zooplankton and phytoplankton communities, as well as native fish populations. Yellow perch are also a host for viral hemorrhagic septicemia, spring viremia of carp virus, and infectious hematopoietic necrosis (USFWS 2019).

Distribution and Abundance Within the Study Area

Like the largemouth bass, yellow perch have spread throughout the country through intentional stocking for the sport fishery. Yellow perch were introduced to Washington in the 1890s and have since spread to 52 watersheds, including the Deschutes River watershed (Fuller and Neilson 2020b; USGS 2020). The inventory of invertebrates and vertebrates conducted on Capitol Lake documented fewer than 50 yellow perch (Herrera 2004). Yellow perch is not likely significantly impacting native wildlife in the Capitol Lake Basin based on its current abundance.

Management Approaches

Management options for yellow perch populations focus on harvest regulations, similar to small- and largemouth bass management described above. More liberal fishing regulations help control population to reduce predation on native fish species.

4.2.1.3 Waterfowl and Mammal

Canada Goose

The Canada goose (*Branta canadensis*) has a black head and crown, a long black neck, and white cheek patches that connect under the chin. The Canada goose prefers open habitats (mowed lawns, parks, golf courses) that are adjacent to open freshwater or estuarine waters. They are extremely successful at living in human-altered areas and establish breeding colonies in urban and cultivated areas that provide food and few natural predators. Although they are extremely adaptable, freshwater source is a major component to their



Canada goose (WDFW 2020c).

successful establishment of an area. Moderate salinity concentrations without freshwater sources can result in slower development, growth, and saline-induced mortality of the Canada geese.

In parks and shorelines with short grass, large flocks of geese can denude areas of vegetation and litter areas with droppings and feathers. Public swimming areas used by many geese have been closed to swimming due to public health concerns. Canada geese are not considered to be a significant source of any infectious disease transmittable to humans or domestic animals, but their droppings can create water quality problems in municipal lakes and ponds from elevated fecal bacteria and nutrient concentrations.

Distribution and Abundance Within the Study Area

Although the Canada goose is native to North America and breeds in Canada and the Northern United States, they have become a nuisance species in recent years due to the substantial population growth following the removal of natural predators and the abundance of safe, man-made bodies of water near food sources. Contrary to its normal migration routine, large flocks of Canada geese have established permanent residence along the Pacific coast from southwestern British Columbia to northern California. The Canada goose is not likely significantly impacting water quality or native plants and wildlife in the Capitol Lake Basin based on its current abundance.

Management Approaches

No single solution is likely to solve conflicts with geese, but instead an integrated approach using several techniques in combination may be required. Public outreach and interpretative signs warning against feeding geese will force them to relocate when the food supply decreases (WDFW 2020c). Landscaping can be used to transform an open lawn to incorporate more native vegetation that is less desirable to geese, reduce the overall size of the lawn, and/or create plant barriers to discourage geese from landing and nesting (WDFW 2020c). Harassment and scare tactics need constant maintenance to remain effective, but can include flags and streamers, scarecrows, noisemaking devices, dogs, and chemical repellents (taste-aversion products) applied to problem areas (WDFW 2020c).

In 1999, the USDA was contracted to manage geese on Capitol Lake by addling eggs in nests and removing Canada geese from the lake during May and July. In the early 2000s, the CLAMP steering committee had a goal to maintain a population of 100 or fewer Canada geese. To meet the CLAMP goal, the USDA removed 486 geese from the Capitol Lake Basin in 2000 and 393 geese in 2001 (WDGA 2005). The USDA Wildlife Services conducted a limited survey of Capitol Lake in 2017 in conjunction with a nutria survey and observed 142 geese (USDA 2018).

Nutria

Nutria (*Myocastor coypus*) is a Prohibited (WDFW 2020b) and High Priority (WISC 2020c) mammal AIS. Nutria are semiaquatic rodents native to South America. Adults are approximately two feet long with dark brown fur and large, orange teeth. Although they are often mistaken for beavers, nutria have a thin tail. Nutria breed year-round and can produce up to three litters a year, with a litter size ranging from 2 to 9 young. In their introduced range, nutria have few natural predators.



Nutria (WISC 2020b).

Although they are well adapted for movement on land, nutria are more at home in the water and prefer slow-flowing streams, lakes, and freshwater marshes as well as brackish and saltwater habitats. Nutria

are herbivores and feed particularly on wetland plants, targeting the base of plant stems and digging for roots and rhizomes in the winter. They often construct circular platforms of compacted, coarse emergent vegetation for use during feeding, birthing, resting, and grooming. They also construct burrows in levees, dikes, and embankments.

Nutria negatively impact invaded ecosystems. Their feeding activity destroys marsh vegetation, transforming marsh areas into open water and displacing native species (Hilts et al. 2019); their burrows undermine water management infrastructure and destabilize banks, increasing erosion along shorelines (Carter and Leonard 2002); and they host infectious diseases that affect humans, livestock, and wildlife (Woods et al. 1992; Drake 2005).

Distribution and Abundance Within the Study Area

In 1935, nutria were brought to Washington for use in the fur industry. Whether they escaped or were intentionally released when fur farming was no longer profitable, nutria spread rapidly throughout western Washington. Nutria observations in Capitol Lake were first recorded in 1975 (Entranco 1997). Although nutria is a high-priority species based on its potential impacts, it is not likely significantly impacting water quality or native plants and wildlife in the Capitol Lake Basin based on its current abundance.

Management Approaches

Feral populations of nutria are managed by shooting and trapping. Eradication is preferable for small to medium size populations, but some level of control is essential in most cases if eradication is not feasible. Fences, walls, and other structures can reduce nutria damage, but high costs usually limit their use. No chemical repellents for nutria are currently registered. Other rodent repellents (such as Thiram) may repel nutria, but their effectiveness has not been determined (LeBlanc 1994).

The USDA Wildlife Service was under contract from 2014–2019 to manually control the population. In 2017, the USDA Wildlife Service conducted a survey for areas of fresh nutria activity and removed one nutria (USDA 2018). No nutria were observed during night survey efforts, and an estimated number of nutria in the basin was not determined (USDA 2018).

4.2.2 Presence in Upstream Waters

The documented presence of aquatic invasive animal species in upstream waters that flow into Capitol Lake, including the Deschutes River, Percival Creek, and nearby freshwater lakes are described below.

4.2.2.1 Invertebrates

The surveys conducted by Johannes (2011, 2013, and 2015) in Capitol Lake also reviewed the surrounding area to document the spread of New Zealand mudsnail outside the immediate study area. In 2013, the New Zealand mudsnail was found for the first time near the mouth of the Deschutes River (Johannes 2013; USGS 2020).

The European ear snail was not found in any sites surveyed outside Capitol Lake (Johannes 2011; Johannes 2013; USGS 2020).

During surveys of Capitol Lake and the surrounding watersheds, the Asiatic clam was found in Black Lake and at one of its outlets, but no documented sightings in the Percival Creek that flows from Black Lake into Capitol Lake (Johannes 2013; USGS 2020).

4.2.2.2 Fish

There have not been any comprehensive surveys documenting the presence of common carp in and in waters flowing into Capitol Lake. Although common carp are documented in Long Lake (USGS 2020), most information on potential presence is based on literature reviews (Herrera 2004; Hayes et al. 2008).

The brown bullhead is documented in several waters surrounding Capitol Lake, including the Grass Lake Refuge, Black Lake, Chambers Lake, Hicks Lake, Long Lake, and Southwick Lake (USGS 2020). Black Lake flows through Black Lake Ditch and Percival Creek to Capitol Lake.

Several lakes surrounding Capitol Lake have documented presence of largemouth bass, including Black Lake, Trosper Lake, Barnes Lake, Susan Lake, Munn Lake, Hewill Lake, Ward Lake, Smith Lake, and Chambers Lake (USGS 2020). Some of these lakes are small and may be intentionally stocked for sport fishing.

Yellow perch are found in several lakes and streams surrounding Capitol Lake, including Black Lake, Ken Lake, Grass Lake Refuge, Trosper Lake, Hewitt Lake, Smith Lake, Chambers Lake, Long Lake, Hicks Lake, and Bigelow Lake (USGS 2020). Some of these lakes are small and may be intentionally stocked for sport fishing.

4.2.2.3 Waterfowl and Mammal

The USGS NAS database does not include Canada geese or other waterfowl.

Outside Capitol Lake, nutria is also found above the mouth of the Deschutes River upstream of the South Basin, but not above Tumwater Falls (USGS 2020).

4.2.3 Presence in Budd Inlet

Invasive species are problematic for urban estuaries (Simenstad et al. 2005). In addition to the known AIS in the freshwater habitat, the Capitol Lake – Deschutes Estuary would be susceptible to invasion by nonnative marine species found in Puget Sound estuaries under the Estuary and Hybrid Alternatives and are described here. These species of concern include two invasive crabs, the European green crab (*Carcinus maenas*) and Chinese mitten crab (*Eriocheir sinensis*), which have yet to invade Puget Sound but are prone to inhabit several estuaries from California to British Columbia. Other high-



European Green Crab (UW 2020)

priority invasive marine animal species for Washington state include the Asian marine clam (*Corbula amurensis*), tunicates (sea squirts), and Atlantic salmon (*Salmo salar*) (University of Washington 2020). Three invasive tunicate species are present in Puget Sound, while the Asian marine clam has not been observed in Washington and Atlantic salmon have not been observed in significant numbers in Puget Sound. California sea lion is not a high-priority invasive marine animal species in Washington state (University of Washington 2020).



5.0 Impacts and Mitigation Measures

5.1 OVERVIEW

This section describes the probable significant impacts related to aquatic invasive plants and animals from the No Action Alternative (Section 5.2), Managed Lake Alternative (Section 5.4), Estuary Alternative (Section 5.5), and Hybrid Alternative (Section 5.6). This section also identifies mitigation measures that could avoid, minimize, or reduce the identified impact below the level of significance.

5.2 NO ACTION ALTERNATIVE

Under the No Action Alternative, Capitol Lake would remain closed to the public and AIS would continue to be managed using containment and other methods aimed at maintaining low population densities. The New Zealand mudsnail population is not likely to substantially increase within the lake or move far outside the lake because it appears not to have done so in at least six years between its introduction in 2009 and the most recent monitoring in 2015, particularly in the absence of public access or meaningful intervention, the populations of other AIS invertebrates, fish, waterfowl, and mammals would be expected to continue to expand at current low rates. Based on this, under the No Action Alternative, the risk for AIS in Capitol Lake to spread to otherwise non-invaded water bodies is expected to be **less-than-significant**.

5.3 IMPACTS COMMON TO ALL BUILD ALTERNATIVES

The build alternatives have in common several adverse impacts and beneficial effects associated with construction and operation. The extent of these impacts on AIS may vary between alternatives and are addressed under the impacts and mitigation described for each alternative section. The goal for AIS management under all build alternatives is to prevent the spread and further distribution of AIS, with purple loosestrife and New Zealand mudsnails being the primary AIS of concern. The eradication of New Zealand mudsnails is assumed to not be feasible under any of the build alternatives regardless of the best management practices and mitigation measures implemented because of their resistance to

extreme environmental factors and treatment, and their ability to reproduce and establish new populations from a single survivor. Eradication of the purple loosestrife is possible, but difficult.

5.3.1 Impacts from Construction

Construction-related impacts common to all build alternatives are associated with dredging and placement or export of dredged sediments.

Prior to construction of all action alternatives, Capitol Lake would be treated to significantly reduce AIS populations within the waterbody and reduce the risk of potential spread once construction activities began. Dredging and other construction activities would occur in in the North and Middle Basins. Most or all dredged material would be used within the basin to create habitat areas; this is a key design element to avoid or minimize the transport of AIS species from the project area.

Best management practices to reduce and contain turbidity during dredging would minimize the potential for substantial transport of invertebrate AIS over the 5th Avenue Dam and into Budd Inlet during construction. Turbidity levels would be less than existing conditions during large storm events so it is reasonable to conclude that plant fragments and invertebrate AIS would be contained through construction.

Dredging and placement of dredged material for habitat areas would occur where few Eurasian watermilfoil plants are located and would not likely affect the abundance of this species in the lake because most of the population is located along the southern shorelines of the Middle Basin and within the South Basin, which is largely outside the construction area. Similarly, purple loosestrife, yellow flag iris, and fragrant water lily are generally not located in the areas of dredging, constructed habitat, or new overwater structures. Based on pre-treatment of AIS throughout Capitol Lake, implementation of BMPs to reduce turbidity, and the small portion of the populations located within the construction areas, initial dredging of any build alternative would have **less-than-significant adverse impacts** on AIS populations and distribution. Dredging and placement of dredge materials in the habitat areas may have **minor beneficial effects** due to removal and burial of some plant and invertebrate AIS.

Some dredge material would be transported out of the study area for the Estuary and Hybrid Alternatives, but export is not assumed under the Managed Lake Alternative. Thus, dredge sediment export would provide a potential vector for transmission of purple loosestrife seeds and invertebrate AIS outside the Capitol Lake Basin for two of the three alternatives. However, sediments exported from the Capitol Lake Basin during construction would be treated prior to transport. Treatment methods may include chemical (e.g., salt or Bayluscide) or physical (e.g., desiccation, heating or freezing) techniques that would need to be proven effective prior to transport. The sediment would only be disposed of at an approved upland site; the upland placement site may be monitored to ensure no purple loosestrife plants or other AIS become established. Therefore, there would be a **less-thansignificant adverse impact** on AIS outside the Capitol Lake Basin from export of sediment dredged during construction. All construction equipment would be appropriately decontaminated before entering and leaving the site to prevent import or export of AIS, consistent with state-required protocol for work in areas of AIS. With these measures, construction would have **less-than-significant adverse impacts** related to changes in abundance and distribution, or potential import or export of AIS.

Construction of the build alternatives would have **no impacts** on fish, waterfowl, and mammal AIS because these animals would avoid construction activities and would not be transported outside the Capitol Lake Basin.

5.3.2 Impacts from Operation

Operation of the build alternatives has a greater potential to impact the distribution and abundance of aquatic invasive plants and animals than the No Action Alternative due to resumption of active recreational use and recurring maintenance dredging.

Active use of the waterbody would be restored under the action alternatives following construction. Long-term operations common to all build alternatives with potential impacts on AIS include:

- Recurring maintenance dredging to maintain target depths.
- Habitat area maintenance (including removing plant AIS).
- Recreational use involving pedestrian and bicycle use of new boardwalks and a dock, and non-motorized watercraft access to the Capitol Lake Deschutes Estuary.

The operation of the build alternatives is not likely to affect the abundance or distribution of aquatic invasive plants in Capitol Lake or other lakes in the study area, provided that the measures outlined in a project-specific AIS Management Plan and BMPs are implemented, including use of decontamination stations, educational signage and ongoing monitoring of AIS. The constructed habitat areas would increase the amount of shallow-water habitat preferred by the New Zealand mudsnail and other invertebrates. After being planted with native emergent, wetland, and upland native species, the habitat areas would increase the amount of forage and refuge habitat for the Canada goose and nutria. The constructed habitat area varies among the build alternatives but would be small relative to the overall project area. Given the small amount of shallow water habitat areas would have a **less-than-significant adverse impact** on AIS abundance and distribution. Within the habitat areas, Enterprise Services would continue to manage aquatic invasive plant species, limiting their expansion. As part of a Habitat Enhancement Plan for the constructed habitat areas, aquatic invasive plant survivability.

The risk of expanding the distribution of AIS from maintenance dredging is considered low because prior to construction, the Capitol Lake Basin would be treated to substantially reduce and/or eradicate plant and invertebrate AIS. The handling of sediment dredged during maintenance dredging varies by alternative and is discussed in more detail below. Maintenance dredging would have **no impact** on the distribution or population size of fish, waterfowl, or mammal AIS.

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Under all action alternatives, portions of the basin would be open to pedestrian traffic along the boardwalks and dock, and to non-motorized boating activity limited to watercraft carried by hand. The risk for importation of new plant AIS or exportation of existing plant AIS from reintroduced boating in the Capitol Lake Basin would be low because boating would be limited to hand-carried watercraft launched from designated boat access locations in the project area, and AIS plants (including fragment and seeds) largely exist in shallow areas along the shorelines, not in the deeper waters where boating would be focused. Plant AIS are primarily imported and exported to water bodies by plant fragments and seeds that collect in motorboats and trailers. Hand-carried nonmotorized watercraft are not likely to transport a substantial amount of plant fragments and seeds. Additionally, a critical component of the action alternatives is the installation and operation of decontamination stations at the designated boat access locations of the build alternatives would have **less-than-significant adverse impacts** on plant AIS because operations are not anticipated to substantially affect the abundance and distribution of invasive plant populations within or outside the study area.

The increase in traffic and activity on the shoreline and in the water would increase opportunity for the New Zealand mudsnail to spread outside Capitol Lake. New Zealand mudsnails can survive for long periods of time on hard material, such as shoes, watercraft, and other recreational equipment.

To avoid and minimize this impact, decontamination stations would be installed, maintained, and

operated for all action alternatives at a new boat launch in Marathon Park, the existing boat launch at Tumwater Historical Park and at the Interpretive Park for decontaminating footwear, fishing gear, and non-motorized vessels used in Capitol Lake. A decontamination station may also be installed at the West Bay Park boat launch, if needed, under the Estuary and Hybrid Alternative given the ability for boats to travel from Budd Inlet to the Capitol Lake Basin once the 5th Avenue Dam is removed. Decontamination of vessels and gear is highly effective in stopping the spread of New Zealand mudsnail to other waterbodies.



Lake Whatcom Decontamination Station (T. Ward)

The build alternatives would also include educational signs that warn recreational lake users of the presence of New Zealand mudsnails and their potential to spread. Signage would notify recreationalists that water access is only permitted in areas where a decontamination station is provided; this helps to ensure decontamination is occurring. While the educational signs would not entirely prevent further spread of New Zealand mudsnails, they would inform the public of the importance and requirement of using the decontamination stations.

The decontamination stations initially would be attended to educate users and ensure 100 percent compliance. Attended stations would be staffed by trained personnel to provide inspection and

decontamination of all watercraft and personal equipment. It is anticipated that the stations would be attended during daylight hours every day of the week except holidays. Monitoring would be conducted to confirm and track the use and effectiveness of attended stations. If monitoring indicates recreationalists are effectively using the stations and very few AIS are present on equipment, then the stations may be converted to unattended stations in the future. Effective use of education and decontamination stations is considered necessary to reduce impacts to **less-than-significant** levels for the spread of AIS.

5.4 MANAGED LAKE ALTERNATIVE

5.4.1 Impacts from Construction

In addition to impacts common to all build alternatives, construction impacts of the Managed Lake Alternative on AIS would primarily be associated with the following:

- Dredging in the North Basin and using dredge material to create habitat areas. Maintaining dredged material within the system is a key design element that avoids the transport of AIS species from the project area.
- Repairing the 5th Avenue Dam

Impacts from initial dredging and other construction activities would be as described in Section 5.3.1. Implementation of best management practices during construction would limit the potential export or import of plant and invertebrate AIS resulting in **less-than-significant impacts** and some **minor beneficial effects** related to potentially burying aquatic invasive plants and New Zealand mudsnails. There would be **no impacts** to AIS associated with repairing the 5th Avenue Dam because all repair work would be contained with the spillways, conducted overwater, or conducted on the Budd Inlet side of the dam where no known invasive species are present. Construction of the Managed Lake Alternative would have **no impacts** on fish, waterfowl, or mammal AIS.

5.4.2 Impacts from Operation

Operational impacts of the Managed Lake Alternative on AIS would generally be as described in Section 5.3.2, impacts common to all alternatives. Active use of the project area would be restored following construction with impacts primarily associated with the following:

- Recurring maintenance dredging in the North Basin to maintain target depths
- Pedestrian and bicycle use on boardwalks
- Non-motorized boating in the lake
- Fishing from a reconstructed dock and boats

Operation of the Managed Lake Alternative would have **less-than-significant adverse impacts** on plant and animal AIS because operations are not anticipated to substantially affect the abundance and distribution of invasive plant and animal populations within or outside the study area.

Under the Managed Lake Alternative, maintenance dredging would occur in the North Basin. The sediment that is exported after maintenance dredge events would be chemically or physically treated, as required by AIS transportation regulations, to prevent the transport of live New Zealand mudsnails. After treatment and during transport, sediments would be covered and only disposed of at an approved upland site. The upland placement site may be monitored to ensure no AIS become established at the placement site.

Material dredged from the Managed Lake would not be suitable for placement at an open water disposal site in Puget Sound because of the presence of the New Zealand mudsnail, which is not expected to be eradicated from the freshwater environment, and would persist at the high densities similar to those under the No Action Alternative. Best management practices and compliance with AIS transportation regulations would result in **less-than-significant adverse impacts** related to changes in abundance and distribution of New Zealand mudsnails and other invertebrate AIS. Maintenance dredging activities would have **no impact** on the distribution or population size of fish, waterfowl, or mammal AIS.

Portions of the lake would be open to pedestrian traffic along the boardwalks and dock, and to nonmotorized boating activity limited to watercraft carried by hand. The impacts associated with increased recreational use would be as described in Section 5.3.2. Effective use of educational signage and decontamination stations would result in **less-than-significant** adverse impacts on the spread of New Zealand mudsnails and other invertebrate AIS to other freshwater bodies. There would be the opportunity to introduce AIS to Capitol Lake from outside sources, but the potential for new AIS to be introduced to Capitol Lake would be minimized by using decontamination stations upon both entry and exit from the lake.

The reintroduction of fishing within Capitol Lake would have **minor beneficial effects** by reducing invasive fish species. This management approach is effectively used in other systems for controlling AIS populations. There would be **no impact** on waterfowl or mammal AIS.

5.5 ESTUARY ALTERNATIVE

5.5.1 Impacts from Construction

Construction impacts of the Estuary Alternative on AIS would generally be as described in Section 5.3.1, impacts common to all alternatives. Construction impacts on AIS associated with the Estuary Alternative primarily relate to the following activities:

- Dredging in the North Basin and Middle Basin and using dredged material to create habitat areas in both basins. Maintaining dredged material within the system is a key design element that avoids or minimizes the transport of AIS species from the project area.
- Removing the 5th Avenue Dam following the completion of dredging and habitat area construction, allowing the entire Capitol Lake Basin to become filled with marine water, creating a brackish estuarine environment.

In addition to impacts common to all build alternatives, construction impacts of the Estuary Alternative on aquatic invasive plants and animals would primarily be associated with removal of the 5th Avenue Dam. Implementation of best management practices and decontamination during dredging and other construction activities would limit the potential export or import of plant and invertebrate AIS resulting in **less-than-significant adverse impacts**. Best management practices to reduce and contain turbidity during dredging would result in **less-than-significant adverse impacts** as actions are not anticipated to significantly increase transport of AIS outside the Capitol Lake Basin. Removal and burial of a small proportion of the plant and invertebrate AIS populations would result in **minor beneficial effects**.

Although the Capitol Lake Basin would be treated prior to construction, removing the 5th Avenue Dam could have **significant adverse impacts** from the potential transport of purple loosestrife, if any plant fragments or seeds remained viable after treatment. Purple loosestrife seeds are salt tolerant and could become plants if they settle near a freshwater stream or river mouth. Impacts would be avoided if the purple loosestrife population in Capitol Lake is eradicated or dramatically reduced several years before dam removal to eliminate viable seeds in sediments (see mitigation measures described in Section 5.7). Dam removal would have **less-than-significant adverse impact** on other invasive plant species in Capitol Lake because these species have a general lack of salt tolerance so they are unlikely to successfully travel and become established in a nearby freshwater stream or river; or if they do become established, they are unlikely to impact native species due to their common presence in the region.

Dam removal would also flush a significant number of New Zealand mudsnails into Budd Inlet. There would be a high mortality from this initial flush. However, some New Zealand mudsnails could potentially drift to freshwater streams and rivers discharging to the Puget Sound shoreline. New Zealand mudsnails and other invertebrate AIS are somewhat tolerant of saltwater and could survive several days in tidal currents, especially if they are transported on floating debris. The potential spread of New Zealand mudsnails outside of the project area would be considered a **significant impact**. This impact could be minimized by treating the population in Capitol Lake prior to removing the dam, as described further as mitigation measures in Section 5.7.

Construction of the Estuary Alternatives would have **no impacts** on fish, waterfowl, and mammal AIS because these animals would avoid construction activities and would not be transported outside the Capitol Lake Basin. Fish AIS would be transported outside the basin during dam removal, but are unlikely to survive transport in marine waters to become established in a stream or river.

5.5.2 Impacts from Operation

Operational impacts of the Estuary Alternative on AIS would generally be as described in Section 5.3.2, impacts common to all alternatives. Active use of the project area would be restored following construction with impacts on AIS primarily associated with the following:

- Recurring maintenance dredging would occur in impacted areas of West Bay.
- Pedestrian and bicycle use of boardwalks along the shoreline.

- Access to the estuary would be open to recreational non-motorized watercraft and decontamination stations would be provided in Marathon Park, Tumwater Historical Park Interpretive Park, and potentially West Bay Park.
- Incidental use of the estuary would also occur via boat access from Budd Inlet.

5.5.2.1 Aquatic Invasive Plants

The impact of recreational access on aquatic invasive plants under the Estuary Alternative would be as described in Section 5.3.2, impacts common to all alternatives. Enterprise Services would continue to manage aquatic invasive plant species, limiting their expansion.

Removal of the 5th Avenue Dam could increase the long-term movement of seeds and plant fragments into Budd Inlet downstream of the study area. However, transport of AIS occurs under existing conditions when sediment and other debris is discharged during high river flow events. These events are defined as levels of 11 feet or more on the Deschutes River near the Rainier Station, and were examined on the National Weather Service, Advanced Hydrologic Prediction Service (NOAA 2020). In the past 20 years, the Deschutes River exceeded flood stage 25 times at an average frequency of around once per year (NOAA 2020). During these events, seeds, plant fragments and other debris are moved over the 5th Avenue Dam and into West Bay. And despite this transport, the population and distribution of AIS have not measurably increased outside of the project area. Comparatively, under the Estuary Alternative, there will be fewer freshwater plant AIS populations due to saltwater affecting their abundance. Therefore, the introduction of saline waters in the Estuary Alternative would likely have **substantial beneficial effects** related to decreased distribution and abundance of freshwater plant AIS, primarily those saltwater-intolerant species other than purple loosestrife.

Purple loosestrife is the only freshwater plant AIS within the Capitol Lake Basin that is tolerant of saltwater. Treatment would occur prior to construction in an attempt to eradicate the species, but some plant fragments or seeds may still exist. Thus, purple loosestrife may not be eradicated when saline waters enter the Capitol Lake Basin after dam removal. Other plant species have low to no tolerance to brackish water and would likely be eliminated in the North and Middle Basins, but these species may survive the low salinities expected in the South Basin, Percival Cove, and the Interpretive Center ponds. Enterprise Services would continue to manage the remaining aquatic invasive plant species, limiting their abundance and distribution. The risk of new plant AIS invading the estuary would be low due to saline conditions and boat launching being limited to hand carried watercraft; and motorboats from marine waters are not likely to carry marine or freshwater AIS plants and would not travel close to vegetated areas of the estuary shoreline.

The risk for importation of new plant AIS or exportation of existing plant AIS from reintroduced boating in the Capitol Lake Basin would be low because boat launching would be limited to hand-carried watercraft at designated boat access locations. Decontamination stations would also be provided and staffed at these locations as described in Section 5.3.2. Also, if incidental motorboat or non-motorized watercraft access occurs via Budd Inlet to the Estuary Alternative, the vessels would have limited contact with plant fragments and seeds that exist along the shoreline in shallow water, which is not

conducive to boating. Thus, recreational access would result in **less-than-significant adverse impacts** to plant AIS.

Maintenance dredging of impacted areas of West Bay would have **no impact** on plant AIS because none would be present in the saline waters.

5.5.2.2 Aquatic Invasive Animals

The impact of recreational access on aquatic invasive animals under the Estuary Alternative would be as described in Section 5.3.2, impacts common to all alternatives.

Following the removal of the dam and the initial flush, several freshwater aquatic invasive animals that are tolerant of brackish water would continue to be present near freshwater sources, although with much more limited distribution and abundance. In particular, New Zealand mudsnails acclimatized to increased salinity can survive for long periods in the salinities observed in Budd Inlet. Removal of the dam barrier would increase the potential for suspended New Zealand mudsnails, either individually or attached to debris, to be washed into Budd Inlet by high river flow during low tides.

Exported New Zealand mudsnails could settle at the shallow mouths of streams feeding Budd Inlet, and possibly beyond. However, there is no indication that these streams have become infested since the New Zealand mudsnails invaded the lake in 2009 and, presumably, they have been discharged through the dam in high river flows. Although river discharge rates would not increase with the Estuary Alternative, sediment and debris input to Budd Inlet would increase without the 5th Avenue Dam. Although mudsnail-specific monitoring has not been conducted since 2015, the mudsnail is not likely to have spread much since 2015 because its presence has not been reported in nearby streams or estuarine waters.

The mudsnail population may increase over time in the Capitol Lake Basin as they become acclimatized to estuarine waters under the Estuary Alternative, which has occurred in the Columbia River Estuary (Hoy et al. 2012). However, it is likely that establishment of the New Zealand mudsnail along the shorelines of Budd Inlet would not have a significant impact on native aquatic species because there is limited evidence that the New Zealand mudsnail had ecosystem level impacts or cascading ecological effects since its invasion of the Pacific Northwest over 25 years ago (Draheim 2015). Given the apparent lack of downstream spread over the past 10 years and the significantly decreased abundance expected in the Estuary Alternative, the potential increase in transport and survival outside the study area by the dam removal would have a **less-than-significant impact** on New Zealand mudsnails.

Compared to the Managed Lake Alternative, the distribution of the New Zealand mudsnail may be wider, but the population is expected to be significantly reduced. The New Zealand mudsnail does not thrive in a saltwater environment, and as described above, there is no existing evidence of significant ecological harm from low density populations of New Zealand mudsnails.

As described in Section 5.3.2, the increase in pedestrian and non-motorized watercraft use would increase the potential for spread of invertebrate AIS outside the study area. Although the New Zealand

mudsnail population would be significantly reduced by the conversion to a brackish environment, the increase in activity on and around the estuary would potentially increase spread of New Zealand mudsnails to other freshwater bodies by equipment (boots and boats) contacting estuary sediments. Any incidental motorboat access from Budd Inlet is not likely to import or export invertebrate AIS from the estuary because they are not likely to contain invertebrate AIS upon entering the estuary or to contact nearshore sediments where the AIS may continue to be present within the estuary. The implementation of educational signs and decontamination stations described in Section 5.3.2.2 would dramatically reduce the potential spread to a **less-than-significant adverse impact** on invertebrate AIS.

Although most of the fish AIS are somewhat tolerant of brackish water found in estuaries, few, if any, would be expected to establish themselves in a brackish habitat or in other freshwater habitats along the Puget Sound shoreline. Therefore, removing the 5th Avenue Dam would have a **less-than-significant impact** on fish AIS because there would not be a substantial increase in abundance or distribution outside of the study area.

The Canada goose is also tolerant of brackish and saltwater, but this waterfowl AIS may experience slower development, growth, and saline-induced mortality without a large source of freshwater. Unlike the invertebrate and fish AIS, the Canada goose distribution is not currently limited by the 5th Avenue Dam. Removing the 5th Avenue Dam would have **minor beneficial effects** related to restricting Canada geese to areas of freshwater inputs.

Nutria are found in brackish and saltwater environments and would tolerate the transition to an estuarine environment. Because their distribution is not limited by the 5th Avenue Dam, the dam removal would have **no impact** on nutria.

Recreational use of the estuary would have **no impact** on any remaining fish, waterfowl, or mammal AIS populations.

Maintenance dredging under the Estuary Alternative would occur in impacted areas of West Bay only, not within the Capitol Lake Basin. The New Zealand mudsnail is not expected to be in the sediment that would be dredged under the Estuary Alternative because of the salinity levels within West Bay and because maintenance dredging would occur in deeper water used for navigation. In addition, the dredged sediments would have primarily originated from the Deschutes River, which does not have an established population of New Zealand mudsnails.

Although New Zealand mudsnails are tolerant of higher salinity levels that can be found in West Bay, few if any are assumed to be present at the target dredging depths because the New Zealand mudsnail prefers shallow water habitat. For this reason, sediment dredged during the maintenance dredging could be suitable for placement at an open water disposal site in Puget Sound. The sediment would have to be sampled for New Zealand mudsnails and purple loosestrife seeds to demonstrate suitability, and the environmental agencies with jurisdiction would have to be confirmed, refer to the *Sediment Quality Discipline Report* for more detail.) In-water placement of dredged sediment would result in a significant

costs savings for the project compared to upland disposal; it also reduces truck trips from surface streets and the associated greenhouse gas emissions.

If the sediment is not suitable for in-water placement, it would be transloaded into trucks and hauled to an upland placement site, using BMPs consistent with those described for the Managed Lake Alternative.

Maintenance dredging activities would have **no impact** on distribution or abundance of invertebrate, fish, waterfowl, or mammal AIS because no animal AIS would likely be present in the dredging area.

5.6 HYBRID ALTERNATIVE

5.6.1 Impacts from Construction

Construction impacts of the Hybrid Alternative on AIS would generally be as described for the Estuary Alternative in Section 5.5.1. Construction impacts on AIS associated with the Hybrid Alternative primarily relate to the following activities:

- Dredging in the North Basin and Middle Basin and using dredge material to create habitat areas. Maintaining dredged material within the system is a key design element that avoids or minimizes the transport of AIS species from the project area.
- Removing the 5th Avenue Dam following the completion of dredging and habitat area construction, allowing the entire Capitol Lake area to become filled with saltwater, creating a brackish estuarine environment and saltwater reflecting pool.

The impact of construction activities on aquatic invasive plants and animals under the Hybrid Alternative would be as described for the Estuary Alternative in Section 5.5.1. For most plant and invertebrate AIS, construction activities would have **less-than-significant impacts** or **minor beneficial effects**. Although the Capitol Lake Basin would be treated prior to construction, removing the 5th Avenue Dam could have **significant adverse impacts** on purple loosestrife, if any plant fragments or seeds remained viable after treatment. There would be **no impacts** on plant AIS from constructing the reflecting pool.

Dam removal could also flush a significant number of New Zealand mudsnails into Budd Inlet that could potentially drift to freshwater streams and rivers discharging to the Puget Sound shoreline. The potential spread of New Zealand mudsnails outside of the project area would be considered a **significant impact**. Construction activities would have **no impacts** on fish, waterfowl, and mammal AIS.

5.6.2 Impacts from Operation

Operational impacts of the Hybrid Alternative on AIS would be as described for the Estuary Alternative in Section 5.5.2. Active use of the project area would be restored following construction with impacts on AIS primarily associated with the following:

• Recurring maintenance dredging would occur in impacted areas of West Bay.

- Pedestrian and bicycle use of boardwalks along the shoreline.
- Access to the estuary would be open non-motorized watercraft and decontamination stations would be provided in Marathon Park, Tumwater Historical Park and Interpretive Park, and potentially West Bay Park.
- Incidental use of the estuary would also occur via boat access from Budd Inlet.

The impacts associated with the operation of the Hybrid Alternative would be as described for the Estuary Alternative in Section 5.5.2. Recreational access would result in **less-than-significant adverse impacts** to plant AIS. The introduction of saline waters in the Hybrid Alternative would likely have **substantial beneficial effects** on the distribution and abundance of freshwater plant AIS, primarily those saltwater intolerant species other than purple loosestrife.

The potential increase in downstream transport of invertebrate AIS outside the study area from removal of the 5th Avenue Dam would have a **less-than-significant impact** on New Zealand mudsnails due to the reduced population expected in the estuarine waters. Dam removal would have **less-than-significant impacts** on fish AIS, **minor beneficial effects** on Canada geese, and **no impact** on nutria.

The implementation of educational signs and decontamination stations would dramatically reduce the potential spread to a **less-than-significant adverse impact** on invertebrate AIS. Recreational use of the estuary would have **no impact** on any remaining fish, waterfowl, or mammal AIS populations.

Maintenance dredging of impacted areas of West Bay would **no impact** on plant AIS because none would be present. Maintenance dredging activities would have **no impact** on distribution or abundance of invertebrate, fish, waterfowl, or mammal AIS because no animal AIS would likely be present in the dredging area.

5.7 AVOIDANCE, MINIMIZATION, AND MITIGATION MEASURES

Enterprise Services would avoid and minimize potential impacts by complying with regulations, permits, plans, and authorizations. These anticipated measures, and other mitigation measures that could be recommended or required, are described below.

5.7.1 Measures Common to All Build Alternatives

5.7.1.1 AIS Management Plan

An AIS adaptive management plan would be developed and implemented for the preferred alternative during the future phase. Elements of the adaptive management plan would vary depending on the alternative and AIS abundance, and generally include:

• Conduct monitoring of New Zealand mudsnails to identify their abundance within Capitol Lake and adjacent waters and of purple loosestrife seeds to identify their abundance and viability in lake sediments.

- Determine which chemical treatment tests should be conducted and can be permitted by Ecology's National Pollutant Discharge Elimination System (NPDES) permit program and the associated treatment restrictions for reducing the New Zealand mudsnail population before construction to reduce its potential spread during construction.
- Design and conduct New Zealand mudsnail treatment tests with chemicals known to be effective by application to lake waters or exposed sediments during a drawdown, and with site-specific environmental controls using saltwater backflushing and drawdown desiccation without chemicals.
- Obtain experimental use authorization to apply and test effectiveness of select chemicals that are not included in the Aquatic Invasive Species Management Permit (Ecology 2020b), and add effective chemicals to a future permit update for site-wide use prior to construction.
- Prepare and implement a New Zealand mudsnail treatment plan using the preferred methodology if and where the population is large enough to warrant treatment.
- Prepare and implement a purple loosestrife treatment plan with a goal of eradication before construction begins to avoid or minimize downstream migration of seeds during operations.
- Specify best management practices for avoiding or minimizing the export of AIS through the dam during construction, such as the use of turbidity curtains and AIS monitoring.
- Conduct long-term monitoring of New Zealand mudsnails and purple loosestrife in the study area and adjacent waters to track changes in abundance for adaptive management.
- Research and design attended or unattended decontamination stations and establish a maintenance and monitoring plan to ensure their continued effectiveness.
- Design and install education signs at strategic locations to inform citizens of the AIS threat and requirements for preventing their spread.

5.7.1.2 During Construction

No transport of High Priority AIS outside of Capitol Lake is allowed by state law RCW 77.135. High Priority AIS in the Capitol Lake Basin include purple loosestrife, Eurasian watermilfoil, New Zealand mudsnail, and nutria. The risk for AIS transport would be reduced to low by implementing the following best management practices during construction:

- Vessels entering and leaving Capitol Lake would meet "clean, drain and dry" requirements under RCW 77.135.110 as described by WDFW (2016). These requirements include cleaning equipment that has come into contact with Washington waters by Level 1 and 2 decontamination protocols (WDFW 2016). Level 1 protocols require removing any visible native and nonnative plants, algae, or mud from boat hulls and engines, trailers, and other gear; draining any accumulated water from boats or gear, including water used in cleaning, back into the water body from which it came; rinsing all surfaces with potable water; and letting boats or gear fully dry for 48 hours before using again in Washington waters.
- Vessels and construction equipment used near the ordinary high-water mark and personal protective equipment such as boots also would be decontaminated by Level 2 protocols

before entering and leaving the study area. Level 2 protocols are required for high-risk situations and include treatment with either Virkon Aquatic[®], hot water, freezing, or Formula 409[®] (WDFW 2016).

• Operators of vessels and construction equipment to be used in Capitol Lake would obtain an Aquatic Invasive Species Prevention Permit from WDFW to educate themselves upon the presence of AIS and how to prevent the spread of AIS prior to work in the study area per RCW 77.135.210. Most to all dredge material from construction would remain within the system to further avoid potential spread of AIS. Dredge material exported for the Estuary and Hybrid Alternatives would be treated to remove viable AIS before transportation to an approved upland location as described for maintenance dredging during operation.

5.7.1.3 During Operation

During maintenance dredging, vessels and dredging equipment would follow the same regulatory requirements described for construction above. All transported dredge material would follow the AIS management plan for transport and any upland disposal of dredge materials containing viable purple loosestrife seeds or New Zealand mudsnails, which includes:

- Treating materials in trucks or railcars to kill New Zealand mudsnails.
- Covering dredge materials to prevent loss of viable seeds or mudsnails during transport.
- Covering dredged material at an upland beneficial reuse or disposal site with a soil layer, not disturbing the disposed materials for a specified period.
- Post-placement monitoring of the upland beneficial reuse or disposal site for a specified duration to ensure no plant growth at the site.
- Conduct sampling of material dredged from West Bay under the Estuary and Hybrid Alternatives to determine whether high priority plant or invertebrate AIS exist within the sediment. If sampling demonstrates that the material is free of AIS, and is chemically suitable, it would be disposed of at an open water disposal site in Puget Sound.

Non-motorized boat access would be restricted to permanent inspection and decontamination stations located in Marathon Park, Tumwater Historical Park, and Interpretive Center for all build alternatives. A decontamination station may also be operated at West Bay Park under the Estuary and Hybrid Alternatives, if needed. All boats leaving the study area would be decontaminated to prevent spread of AlS from the area, but would not be required to obtain an Aquatic Species Prevention Permit.

5.7.2 Measures Specific to Each Build Alternative

5.7.2.1 Managed Lake

The following measures could reduce impacts related to the spread of AIS. These measures could be implemented in addition to those that would be included in regulatory authorizations.

Construction

As part of the AIS Management Plan, a New Zealand mudsnail treatment plan would be prepared with measures to significantly reduce the population prior to and perhaps during construction. This is a critical measure to avoiding or minimizing the spread of AIS during and after construction. The plan would consider the management approaches effective in managing mudsnails within Capitol Lake, which likely include backflushing, drawdown, freezing, desiccation, and/or chemical treatment. The plan would propose the appropriate amount and duration of treatment to reduce the mudsnail population and obtain necessary authorizations for the treatment. An assessment of conditions within the study area would determine the mudsnail abundance and distribution and the frequency of treatment necessary to maintain low abundance. It would also outline measures to avoid the spread of any AIS outside of the study area, such as use of turbidity curtains to contain suspended sediment and debris that may contain AIS and monitoring of discharged waters for the presence of AIS.

Operation

Potential treatment options outlined in the AIS Management Plan could be used to control New Zealand mudsnail abundance and distribution around Capitol Lake if the population dramatically increases over the long term. Depending on efficacy during construction and permitting requirements, chemical treatments could continue to be applied following construction in coordination with the regulatory agencies. The chemical treatments may only target High Priority AIS such as New Zealand mudsnail, purple loosestrife, and Eurasian watermilfoil. Chemical treatments, including sodium chloride and Bayluscide, would be assessed based on New Zealand mudsnail distribution and density, targeting areas of the highest density. However, chemical treatments severely impact native species. It may be difficult to obtain a permit for large-scale chemical treatments of the lake; it is likely to require coordination with the Washington State Departments or Ecology and Fish and Wildlife and other regulatory agencies to authorize use of chemicals that are registered by USEPA but not specifically approved by the Aquatic Invasive Species Management Permit. Thus, the benefits and impacts of treatment would be carefully weighed during preparation of the AIS Management Plan.

While eradication is generally considered not to be feasible given the extent of the New Zealand mudsnail infestation and their resiliency, chemical treatment is a useful method for significantly reducing the population and limiting its spread outside the study area. It would also continue to avoid and minimize potential impacts from AIS to ecological functions and recreation within the project area. Periodic monitoring around sites of known presence and absence of New Zealand mudsnail populations would be used to ensure effectiveness of treatment and decontamination procedures.

5.7.2.2 Estuary and Hybrid Alternatives

The following measures could reduce impacts related to the spread of AIS under both the Estuary and Hybrid Alternatives. These measures could be implemented in addition to those that would be included in regulatory authorizations.

During Construction

As part of the AIS Management Plan, Enterprise Services would prepare a treatment plan to significantly reduce AIS populations prior to and perhaps periodically during construction. This is a critical measure to avoiding or minimizing the spread of AIS during and after construction. The plan would consider backflushing, drawdown, and/or chemical treatment prior to construction and determine the amount and duration of treatment to reduce potential downstream transport from the study area. This treatment would likely impact freshwater native species; however, these species are not likely to survive under the Estuary and Hybrid Alternatives because of increased salinity and continued inflow of marine water during high tide.

Treatment would likely be performed for only purple loosestrife and New Zealand mudsnail due to their high saltwater tolerance and higher risk for infesting downstream waters. Purple loosestrife would be treated multiple times before construction to minimize the viable seed bank in the lake sediments. New Zealand mudsnails may require treatment on only one occasion after construction of the habitat areas in case of contact with exposed organisms, and sufficient time should be allowed for chemical inactivation before dam removal. An assessment of conditions within the study area would determine the frequency necessary to maintain low distribution and abundance and limit the risk of spread outside the study area.

During Operation

Much of the New Zealand mudsnail and other AIS populations would be managed by the introduction of saltwater to the study area after removing the 5th Avenue Dam. Although many species are tolerant to brackish water, the estuarine habitat would significantly reduce their density and limit their distribution to freshwater sources within the study area, which would occur along shallow shoreline areas with groundwater springs and in the South Basin. Due to the likely persistence of New Zealand mudsnails in upstream portions of the Estuary and Hybrid Alternatives, additional mitigation measures such as targeted chemical treatments in areas of high concentrations may be needed to prevent potential significant adverse impacts to ecological functions or recreation from their continued presence. However, additional treatments may be severely limited or not permittable under estuarine conditions due to difficulties containing the chemical and in providing a sufficient contact time to be effective. Chemical treatments, including sodium chloride and Bayluscide, would be assessed based on New Zealand mudsnail distribution and density, targeting areas of the highest density. However, chemical treatments severely impact native species and are not currently authorized for use under the Aquatic Invasive Species Management Permit. It may be difficult to obtain a permit for large-scale application of chemical treatments of the estuary or reflecting pool, and would require coordination with the Washington State Departments or Ecology and of Fish and Wildlife to authorize use of chemicals that are registered by USEPA but not specifically approved by the Aquatic Invasive Species Management Permit. Thus, the benefits and impacts of treatment would be carefully weighed during preparation of the AIS Management Plan.

In addition to the decontamination stations in Marathon Park, Tumwater Historical Park and Interpretive Center for all build alternatives, a decontamination station could also be established within the City-owned West Bay Park for the Estuary and Hybrid Alternatives. Boat and equipment decontamination at this third location would help to minimize the potential spread of high-priority species from boaters that may take out at this downstream location, after boating in the Capitol Lake Basin.

Restoring access to Budd Inlet through removing the 5th Avenue Dam could also provide a vector for estuarine and saltwater invasive species, such as the European green crab (*Carcinus maenas*), Chinese mitten crab (*Eriocheir sinensis*), the Asian marine clam (*Corbula amurensis*), tunicates (sea squirts), and Atlantic salmon (*Salmo salar*) (University of Washington 2020), to inhabit the project area. The presence of estuarine and saltwater high-priority species in the Estuary and Hybrid Alternatives should be monitored along with other aquatic invasive and nuisance species, so that immediate action can be taken if they are discovered. This expectation would be outlined in the AIS adaptive management plan.

5.7.3 Significant Unavoidable Adverse Impacts

There would be no significant unavoidable adverse impacts related to AIS under any of the build alternatives.



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